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Impact of the Clean Air Act on Air Pollution and Infant Health: Evidence from South Korea

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Impact of the Clean Air Act on Air Pollution and Infant Health: Evidence from South Korea*

This paper examines the extent to which the 2005 Clean Air Act introduced in South Korea affected air pollution and infant health. To identify the causal effect, we exploit the time and geographical variations in the adoption of the Act between 2003 and 2006. During this period, the Clean Air Act indeed significantly reduced air pollutants. For example, the $\text{PM}_{10}$ level was reduced by 9 percent. However, the Act’s impact on infant mortality was not statistically significant.

JEL Classification: I18, K32, Q52

Keywords: air pollution, infant mortality, $\text{PM}_{10}$

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I. Introduction

Air pollution has received considerable attention from researchers, practitioners, and the general public. While a growing number of empirical studies have identified the causal impact of air pollution on health and other outcomes, research examining policy interventions to reduce air pollution, particularly in a non-US setting, has been scarce (Arceo et al., 2016).

This paper examines an environmental protection policy aimed to reduce air pollution in South Korea, and estimates its impact on pollution reduction and infant health. Although South Korea is steadily ranked as one of the world’s upper middle-income countries, its ambient air pollution level is much higher than that of its peers, indeed the worst among the OECD countries (66 percent higher than that of the US in terms of urban-population weighted average of annual PM$_{10}$ in 2010). While responding to public outcry concerning poor air quality has become a major political agenda, there are also strong reservations from local business groups, who oppose tightening environmental laws. These groups rely on the finding that a large portion of air pollution in South Korea is due to pollutants traveling from China, and thus such law enforcement would have little impact on air quality in South Korea (Fifield and Seo, 2017). Given the severity of South Korea’s air pollution and the political disputes with respect to environmental policies, it is important to examine the extent to which a domestic environmental protection policy has an impact on pollution reduction. This paper addresses this need by examining the 2005 Clean Air Act (CAA) in South Korea.

The CAA, announced in June 2004 and implemented in January 2005, tightened pollution monitoring for diesel emission cars. The CAA was imposed only in the areas with high air pollution levels in Seoul and its surrounding areas but not the rest of the
country. Using these geographical and time variations, we identify the causal impact of the law using a Difference-in-Difference (DID) framework. Subsequently, by employing the CAA as an instrument variable (IV) for air pollutant level (PM$_{10}$), we assess the impact of ambient air quality on infant mortality. Our identification crucially relies on the assumption that the districts affected by the CAA and the non-affected districts share the same time trend. Using the pre-treatment periods, we present supporting evidence that this assumption is likely to hold.

Conditional on weather conditions, we find the PM$_{10}$ concentration was reduced due to the CAA. Additional analyses suggest that the positive impacts are larger for the areas in which the initial air pollution levels were high. However, we find that the CAA did not reduce the infant mortality rate significantly.

II. Institutional Background

South Korea has been infamous for its severe air pollution, the worst among the OECD countries. Although not as severe as that in metropolitan areas in developing countries such as Beijing and New Delhi, the air pollution levels in major cities of South Korea exceed the level considered unsafe by World Health Organization (WHO, 2016). To address this serious problem, the South Korean Congress introduced the Clean Air Act in December 2003. Details, including the criteria for selecting target areas, were announced in June 2004 and became effective in January 2005. The CAA targeted high pollution areas in Seoul and its surroundings. Although it aimed to reduce overall levels of air pollutants such as SO$_2$ (Sulfur Oxides), NO$_2$ (Nitrogen Oxides), VOC$_s$ (Volatile Organic Compounds), and O$_3$ (Ozone), its main objective was to reduce PM$_{10}$ (Particulate Matter). Specifically, its target was an ambient concentration of PM$_{10}$ lower than 50µg/m$^3$ by 2011 as an annual average to meet the WHO guideline; this target is often referred to as
Interim Target-2 (IT-2). The CAA consists of various policies with different phases. This paper examines the first one, targeting emissions of diesel cars. This regulation mandates that all diesel automobiles registered in the treated areas need to be regularly checked for emission levels. If a car does not meet the emission criteria, its owner has to install an emission reduction device or change the car’s engine. Otherwise, the government may charge a fine of up to USD 5,000, approximately 2 times the average household income in 2003. See details in Section A of the Supplementary Materials.

III. Data and Sample

We compile a dataset with pollution measures, weather conditions, and mortality rates for South Korea during the period from January 2003 to December 2006. The air quality information is based on the monthly average of the ambient air pollution levels for each monitor, provided by the National Institute of Environment Research. Weather conditions such as wind speed and precipitation are found to greatly affect the amount of air pollutants. Thus, we separately collect weather data measurements from the Korea Meteorological Administration. The mortality information is based on the restricted-use microdata of death and birth records, provided by South Korea’s Statistics Bureau. We classify deaths based on their causes since certain causes are more likely to be affected than others (e.g., Arceo et al., 2016; Chay and Greenstone, 2003). For example, air pollution may be more likely to increase risk of internal deaths (e.g., due to respiratory diseases) than that of external deaths (e.g., due to car accidents). Similarly, among internal deaths, air pollution may be more likely to increase risk of death due to cardiovascular-respiratory diseases. For this reason, we first analyze all deaths, narrow

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1 We gathered data from the Korea Meteorological Administration. (2003-2006) “Annual Climatological Report,” which provides average of wind speed, monthly total precipitation, number of days affected by yellow dust, and most frequent wind direction in a month collected from each monitor.
the analysis to only deaths due to internal causes, and then analyze the deaths due to cardiovascular-respiratory diseases.\(^2\) We merge these three datasets by aggregating up to district by month levels. Thus, the unit of observation in our analysis is district by month, a total of 34 districts in a given month, 25 of which were treated by the CAA.\(^3\) Table 1 shows the summary statistics of our sample depending on the stage of the CAA: before the CAA was introduced (column (1)), the period when the CAA was announced but not yet implemented (column (2)), and the period when the CAA was implemented (column (3)).

**IV. Econometric Framework and Identification Strategy**

We examine the difference in air pollution levels between the districts affected by the CAA and those not affected by estimating the following DID specification:

\[
P_{\text{pollutant}}_{d,m,y} = \theta_1 \cdot \mathbb{1}(i \in \text{Treat}; \text{announce}) + \theta_2 \cdot \mathbb{1}(i \in \text{Treat}; \text{implement}) + \omega X_{d,m,y} + \alpha_d + \beta_m + \gamma_y + \epsilon_{d,m,y}\]

where \(P_{\text{pollutant}}_{d,m,y}\) is the level of air pollutant in district \(d\), month \(m\), and year \(y\).

Variables \(\mathbb{1}(i \in \text{Treat}; \text{announce})\) and \(\mathbb{1}(i \in \text{Treat}; \text{implement})\) are the indicators covering July to December 2004, and January 2005 to December 2006, respectively. The former is the period when the CAA was announced to the public but not yet implemented; the latter is the period when the CAA was implemented. By including the former indicator, we allow for the possibility that agents may respond to the CAA in

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\(^2\) Our classification is in line with the International Statistical Classification of Diseases and Related Health Problems (10\(^{th}\) edition, ICD-10) produced by the WHO. See details in Section B of the Supplementary Material.

\(^3\) Although there are 247 districts in South Korea, air pollution monitors covered 127 districts in 2003. Among the 127 districts, we exclude 34 districts located in remote rural areas or islands in South Korea. Finally, among the resulting 93 districts, the Korea Meteorological Administration provides weather condition information only for the 34 districts that we use for our analysis (25 treated and 9 untreated districts). Although eliminating those observations substantially reduces the sample size, our main findings qualitatively remain the same if we use all 93 districts but do not control for the weather variables (see Section C.2 in the Supplementary Material).
expecting its implementation starting from January 2005. Parameters $\alpha_d, \beta_m, \text{ and } \gamma_y$ capture district-, month-, and year- fixed effects. Variable $X_{d,m,y}$ includes district-specific weather conditions: precipitation, wind speed, whether a district is subject to dust blows from China, and dummies indicating main wind direction in a given month.\(^4\) See details in Section B of the Supplementary Materials for construction of variables. Variable $\varepsilon_{d,m,y}$ captures unexplained random shock, clustered at the district level.

Our identifying assumption is that, absent the CAA, the trends in air pollution would have been the same for the districts affected by the CAA and the non-affected districts. To test the plausibility of our assumption, we restrict our sample to January 2003 to June 2004, and estimate a linear regression model including the interaction effects between the two indicators: one for the treated districts and the other for the year 2004. If the two groups of districts share the time trend, then the interaction effects should not be different from 0, which indeed we find. Column (1) of Table 2 (Panel A) reports the estimate of that interaction term, which is not statistically significant at conventional levels.

\textbf{V. Results}

Column (2) of Panel A, Table 2 presents our estimates of the impact of the CAA on air pollution. Conditional on weather conditions, time and district fixed effects, the $\text{PM}_{10}$ concentration was reduced by $3.080 \mu g/m^3$ or 5.07 percent during its announcement period, although the effect just missed standard statistical significance (e.g., P-value is 0.11). During the period when the CAA was implemented, the $\text{PM}_{10}$ concentration was

\(^4\) We include these variables because air pollution levels hinge on the weather conditions and season (Barmpadimos et al., 2011; Lee et al., 2006; Hooyberghs et al., 2005). For example, $\text{PM}_{10}$ formation requires dry and stagnant air (Barmpadimos et al., 2011).
reduced even more - 5.667\mu g/m^3 or 9.34 percent, statistically significant at the 5 percent level.\(^5\)

Next, we examine whether the impact of CAA depends on the initial pollution level. To test this possibility, we interact the two policy variables in equation (1) with a district’s initial PM\(_{10}\) concentration.\(^6\) As shown in column (3) of Table 2, the districts with high initial air pollution show greater reduction due to the CAA in both the announcement and enforcement periods. For the average districts in the treatment group whose initial PM\(_{10}\) is 68.260\mu g/m\(^3\), the CAA reduced the PM\(_{10}\) level by 6.123\mu g/m\(^3\) (8.97 percent) during the announcement period and by 7.918\mu g/m\(^3\) (11.59 percent) during the enforcement period.

Since the CAA may affect other air pollutants as well, we examine the impact of the CAA on NO\(_2\) levels. Column (4) of Panel A in Table 2 suggests that the CAA reduced NO\(_2\) by 1.860ppb during the announcement period and by 1.344ppb during the enforcement period, but neither is statistically significant at conventional levels.

Finally, we examine the extent to which the CAA impacted infant mortality. We regress the number of deaths per 100,000 live births, among those whose age is less than 1 year old in district \(d\), month \(m\), and year \(y\) on the explanatory variables in equation (1).\(^7\) Panel B of Table 2 reports the results. During the announcement period, the CAA reduced infant mortality from all causes and that from internal causes by 8.970 and 8.982, respectively.

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\(^5\) We examine the possibility that the Clean Air Act may affect some districts in control group (externality) but find no strong evidence supporting this possibility. See Section C.3 in the Supplementary Materials.

\(^6\) Specifically, we use the average PM\(_{10}\) level between January 2003 and June 2004 in a district as the district’s initial PM\(_{10}\) level.

\(^7\) This analysis measures the reduced-form effect of the CAA on infant mortality. Note that we decide not to estimate the causal impact of PM\(_{10}\) on infant mortality using the IV approach because the CAA’s impact during the announcement period is not statistically significant, resulting in a weak instrument problem. For the purpose of this study, this reduced-form analysis is suitable because we are primarily interested in the possible impact of CAA on infant mortality, not the causal impact of PM\(_{10}\).
respectively (columns (1) and (2)), although these effects are significant at only the 10 percent level. However, we do not find any statistically significant reduction in infant mortality during the enforcement period or in terms of infant deaths caused by cardio-respiratory causes (column (3)).

IV. Conclusion

This paper examines the impact of South Korea’s 2005 Clean Air Act on pollution levels and infant mortality. Using the DID method, we find that the 2005 CAA substantially reduced PM$_{10}$, and other air pollutants, which lead to a reduction in the infant mortality rate. Our findings suggest that, despite external factors such as pollutants transferred from China, South Korea may benefit from stricter environmental regulations related to improving air quality. Furthermore, even in a developed country with nationwide health care, namely South Korea, air pollution imposes significant health risks to the population, measured by infant mortality, which should not be taken lightly when designing environmental policies.

REFERENCES


Table 1. Summary Statistics

<table>
<thead>
<tr>
<th></th>
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</thead>
<tbody>
<tr>
<td></td>
<td>(1)</td>
<td>(2)</td>
<td>(3)</td>
</tr>
<tr>
<td><strong>Panel A. Treatment Group</strong></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>PM$_{10}$ (µg/m$^3$)</td>
<td>69.166</td>
<td>51.272</td>
<td>59.430</td>
</tr>
<tr>
<td>NO$_2$ (ppb)</td>
<td>37.223</td>
<td>32.869</td>
<td>33.656</td>
</tr>
<tr>
<td>Infant deaths</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>- All causes</td>
<td>37.000</td>
<td>25.449</td>
<td>27.029</td>
</tr>
<tr>
<td>- Internal causes</td>
<td>36.488</td>
<td>25.194</td>
<td>26.788</td>
</tr>
<tr>
<td>- Cardio-Respiratory causes</td>
<td>15.555</td>
<td>13.721</td>
<td>11.202</td>
</tr>
<tr>
<td><strong>Panel B. Control Group</strong></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>PM$_{10}$ (µg/m$^3$)</td>
<td>60.962</td>
<td>49.575</td>
<td>56.126</td>
</tr>
<tr>
<td>NO$_2$ (ppb)</td>
<td>25.795</td>
<td>24.432</td>
<td>24.452</td>
</tr>
<tr>
<td>Infant deaths</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>- All causes</td>
<td>49.728</td>
<td>44.171</td>
<td>34.167</td>
</tr>
<tr>
<td>- Internal causes</td>
<td>48.891</td>
<td>44.171</td>
<td>33.310</td>
</tr>
<tr>
<td>- Cardio-Respiratory causes</td>
<td>17.995</td>
<td>9.396</td>
<td>11.303</td>
</tr>
</tbody>
</table>

*Note:* This table provides mean values of key variables in each period by group. The unit of observations is district by month. The infant deaths are the number of deaths (age less than 1 year old) per 100,000 live births.
Table 2: Impact of the 2005 CAA on Pollution

<table>
<thead>
<tr>
<th>Dep. Model</th>
<th>PM$_{10}$ Identification Test</th>
<th>PM$_{10}$ Treatment Effect</th>
<th>PM$_{10}$ Treatment Effect</th>
<th>NO$_2$ Treatment Effect</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>(1)</td>
<td>(2)</td>
<td>(3)</td>
<td>(4)</td>
</tr>
<tr>
<td>1(i: Treat, Jan-Jun. 2004)</td>
<td>-4.537</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>1(i: Treat, Announce)</td>
<td>-3.080</td>
<td>-1.860</td>
<td></td>
<td></td>
</tr>
<tr>
<td>x Initial PM$_{10}$</td>
<td>(1.887)</td>
<td></td>
<td>-0.090*</td>
<td></td>
</tr>
<tr>
<td></td>
<td></td>
<td></td>
<td>(0.047)</td>
<td></td>
</tr>
<tr>
<td>1(i: Treat, Enforcement)</td>
<td>-5.667**</td>
<td>-1.344</td>
<td></td>
<td></td>
</tr>
<tr>
<td>x Initial PM$_{10}$</td>
<td>(2.370)</td>
<td></td>
<td>-0.116***</td>
<td></td>
</tr>
<tr>
<td></td>
<td></td>
<td></td>
<td>(0.039)</td>
<td></td>
</tr>
<tr>
<td>Obs.</td>
<td>576</td>
<td>1,550</td>
<td>1,550</td>
<td>1,550</td>
</tr>
<tr>
<td>Mean dep.</td>
<td>67.101</td>
<td>60.703</td>
<td>60.703</td>
<td>32.190</td>
</tr>
</tbody>
</table>

Note: The unit of observation is district by month. We include weather controls such as precipitation, wind speed, wind direction, and days of yellow dust. We also control for district, month, and year fixed effects. The unit of PM$_{10}$ is µg/m$^3$ and that of NO$_2$ is ppb (parts per billion). There are 34 districts (25 treated districts and 9 untreated districts). Standard errors are clustered at the district level, reported in parentheses. * Significant at 10%; ** Significant at 5%; *** Significant at 1%.

Panel B. Impact on Infant Mortality

<table>
<thead>
<tr>
<th>Dep. Model</th>
<th>All deaths</th>
<th>Internal causes</th>
<th>Cardio/Respiratory causes</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>(1)</td>
<td>(2)</td>
<td>(3)</td>
</tr>
<tr>
<td>1(i: Treat, Announce)$^a$</td>
<td>-8.970*</td>
<td>-8.982*</td>
<td>-1.111</td>
</tr>
<tr>
<td></td>
<td>(5.029)</td>
<td>(5.009)</td>
<td>(2.627)</td>
</tr>
<tr>
<td>1(i: Treat, Enforcement)$^b$</td>
<td>3.894</td>
<td>4.390</td>
<td>-0.313</td>
</tr>
<tr>
<td></td>
<td>(4.086)</td>
<td>(4.049)</td>
<td>(1.992)</td>
</tr>
<tr>
<td>Mean dep.</td>
<td>33.427</td>
<td>32.976</td>
<td>13.264</td>
</tr>
</tbody>
</table>

Note: The unit of observation is district by month. Each coefficient corresponds to a separate regression. We include weather controls such as precipitation, wind speed, wind direction, and days of yellow dust. We also control for district, month, and year fixed effects. The results reported in the table are weighted by the number of births in the respective cohort. There are 34 districts (25 treated districts and 9 untreated districts). Standard errors are clustered at the district level, reported in parentheses. * Significant at 10%; ** Significant at 5%; *** Significant at 1%.