

Boskovic, Branko; Nøstbakken, Linda

Working Paper

Do Land Markets Anticipate Regulatory Change? Evidence from Canadian Conservation Policy

CESifo Working Paper, No. 5956

Provided in Cooperation with:

Ifo Institute – Leibniz Institute for Economic Research at the University of Munich

Suggested Citation: Boskovic, Branko; Nøstbakken, Linda (2016) : Do Land Markets Anticipate Regulatory Change? Evidence from Canadian Conservation Policy, CESifo Working Paper, No. 5956, Center for Economic Studies and ifo Institute (CESifo), Munich

This Version is available at:

<https://hdl.handle.net/10419/144991>

Standard-Nutzungsbedingungen:

Die Dokumente auf EconStor dürfen zu eigenen wissenschaftlichen Zwecken und zum Privatgebrauch gespeichert und kopiert werden.

Sie dürfen die Dokumente nicht für öffentliche oder kommerzielle Zwecke vervielfältigen, öffentlich ausstellen, öffentlich zugänglich machen, vertreiben oder anderweitig nutzen.

Sofern die Verfasser die Dokumente unter Open-Content-Lizenzen (insbesondere CC-Lizenzen) zur Verfügung gestellt haben sollten, gelten abweichend von diesen Nutzungsbedingungen die in der dort genannten Lizenz gewährten Nutzungsrechte.

Terms of use:

Documents in EconStor may be saved and copied for your personal and scholarly purposes.

You are not to copy documents for public or commercial purposes, to exhibit the documents publicly, to make them publicly available on the internet, or to distribute or otherwise use the documents in public.

If the documents have been made available under an Open Content Licence (especially Creative Commons Licences), you may exercise further usage rights as specified in the indicated licence.



Do Land Markets Anticipate Regulatory Change? Evidence from Canadian Conservation Policy

Branko Bošković
Linda Nøstbakken

CESIFO WORKING PAPER NO. 5956
CATEGORY 12: EMPIRICAL AND THEORETICAL METHODS
JUNE 2016

An electronic version of the paper may be downloaded

- from the SSRN website: www.SSRN.com
- from the RePEc website: www.RePEc.org
- from the CESifo website: www.CESifo-group.org/wp

ISSN 2364-1428

Do Land Markets Anticipate Regulatory Change? Evidence from Canadian Conservation Policy

Abstract

Regulation often evolves, and affected consumers or firms may adjust their behavior in anticipation of potential changes to regulation. Using shifting land use regulation boundaries and oil lease prices from Canada, we estimate the effect of anticipated regulatory change on the value of land. We find that anticipated rezoning decreases the price of unregulated leases. Based on our estimates, not accounting for anticipation underestimates the total cost of the regulation by nearly one-third. Overall, the evidence suggests that anticipation effects are significant and that the cost of anticipated regulation is capitalized into land values.

JEL-Codes: Q580, D440, Q520, Q300.

Keywords: regulation, anticipation, land values, zoning, oil leases, endangered species.

Branko Bošković
Alberta School of Business
University of Alberta
3-23 Business Building
Canada – Edmonton, Alberta T6G 2R6
boskovic@ualberta.ca

Linda Nøstbakken
Norwegian School of Economics
Department of Economics
Helleveien 30
Norway – 5045 Bergen
linda.nostbakken@nhh.no

We are grateful to Saulo Castro for excellent research assistance. We thank Jevan Cherniwchan, Sacha Kapoor, and Patrick Richard for their comments and suggestions, and seminar participants at the 2015 Canadian Resource and Environmental Economics Study Group and the 2015 NHH Workshop on Macroeconomics and Natural Resources. This research was supported in part by the Western Centre for Economic Research and the Environment Canada Economics and Environmental Policy Research Network.

1 Introduction

Land markets are useful for assessing the effect of regulation. In particular, if the cost of complying with regulation is capitalized into land values, one can compare the values of regulated to unregulated land to measure the cost. However, regulation is rarely static and often changes over time, and any anticipated costs from expected future regulation will also be capitalized. Because it is difficult to observe where or when such anticipation arises, it may be challenging to measure separately the effects of existing and anticipated regulation, potentially compromising the evaluation of policy based on land values.

In this paper, we estimate the effects of existing and anticipated land use regulation on the value of land. We base our estimation strategy on a simple model of land use regulation and land prices that allows for the possibility that rezoning may occur in the future, in which case the boundaries defining where the regulation is applied will change. Land markets are competitive, so that the price of a parcel of land is equal to its value. Any factors affecting its value, such as the cost of complying with regulation, are capitalized into its price. If regulation boundaries are anticipated to change, forward-looking markets will incorporate the likelihood of rezoning and capitalize any expected costs of anticipated regulation into the price of affected parcels.

If regulation is applied to parcels in a certain area and regulation boundaries are permanent, one can compare the price of parcels on either side of existing regulation boundaries: parcel values are identical except for their regulation status, so any difference in price is due to the effect of the regulation. If rezoning is anticipated, this strategy no longer works. To see this, suppose regulation boundaries are anticipated to shift out. The information that unregulated parcels near current boundaries will be regulated if rezoning occurs will be incorporated into the value of those parcels. In this case, rather than identifying the cost of regulation, the price difference across existing boundaries underestimates it, and instead identifies the difference between the costs of existing and anticipated regulation. A similar logic applies when boundaries are anticipated to shift inward.

We show how to identify the effects of existing and anticipated regulation on land values if one observes where rezoning is anticipated. Knowing where rezoning is likely to occur allows us to categorize which parcels are anticipated to be regulated in the future. We can then compare the price of regulated and unregulated parcels, holding fixed the anticipated rezoning status, to identify the effect of existing regulation. Comparing the price of parcels that differ only in whether they are likely to be regulated in the future identifies the effect of anticipated regulation.

We apply this empirical approach to data from the Canadian province of Alberta. In particular, we use prices for leases granting the right to develop ‘oil sands’ and land use regulations that aim to protect endangered species from the adverse effects of oil sands production. The regulations impose development restrictions in geographic zones determined as critical habitat for the endangered species, and producers there incur costs of complying with the regulations that they do not incur outside the zones. A key feature of these regulations is that the zones have been gradually expanding over time, and we can observe where rezoning is likely to occur and where it is unlikely to occur. Altogether, this information allows us to categorize leases as being either regulated, currently unregulated but likely to be rezoned in the future, or unregulated now and in the foreseeable future.

Using information on the nearly 3,000 oil sands leases sold between 2003 and 2012 and employing a regression discontinuity design, we find that the price per hectare for a lease inside a regulation zone is 30% lower on average than a lease that is not regulated and is not likely to be regulated in the future. At the mean price per hectare in our sample, this amount is equal to \$238 per hectare in 2012 Canadian dollars. This finding suggests that lease holders do capitalize the cost of regulation into the price they pay for their leases.

For leases that are unregulated but for which rezoning in the future is likely, we find that their prices per hectare are about 16% lower on average than for leases that are unregulated and not at risk of being rezoned. Though this estimate is imprecise, we find this effect depends on the proximity to existing zone boundaries: for leases that are, for example, within 5 kilometers of a zone boundary, the price per hectare discount is 22% and is statistically significant at conventional levels; the effect dissipates the farther is a lease from the existing boundaries. This finding is consistent with the notion that, given the historical practice of gradual zone expansion, leases near existing boundaries are believed to be at greater risk of rezoning than leases farther away. However, realized boundary changes were not gradual but quite expansive, and our results suggest that firms underestimated the likelihood of rezoning of unregulated leases far from existing boundaries. Overall, we take these results as evidence that the costs of anticipated regulation are capitalized into lease prices.

Oil sands lease prices are particularly useful for calculating the total cost of the regulation. The leases are sold in competitive, first-price, sealed-bid auctions. According to auction theory, the winning bid in such auctions is equal to the value of the object, which in this context is the present value (PV) of profits from developing the resource. This implies that the cost of complying with existing or anticipated regulation, which negatively affects profits, should decrease the winning bid by an amount equal to that cost. Based on our estimates,

we find that the total PV cost of the existing and anticipated regulation for leases sold during 2003–2012 is between \$1.6 and \$2.3 billion in 2012 Canadian dollars. The total PV cost based on estimates that fail to account for anticipated regulation underestimates this amount by 29%.

Our study is relevant because anticipated changes to existing regulation, particularly in the context of zoning, are a common feature of policymaking. Our results suggest that land values do fluctuate with anticipation about potential policy changes. Furthermore, the results suggest that failing to account for anticipatory behavior by firms (or consumers) can lead to underestimates of the total cost of regulation, causing problems for evaluating policy in practice. More particular to the context of our application, the behavior caused by anticipated regulatory change has negative implications for broader public policy, since the costs are borne entirely by the government.

To the best of our knowledge, this is the first paper to investigate whether anticipated changes to land use regulation are capitalized into the value of land. A large literature estimates the effect of land use regulation, or zoning, on property values, with a recent example being Turner et al. (2014). We contribute to this literature in two ways. First, we show that when boundaries can change, accounting for anticipation about potential regulatory change is necessary to identify the effect of regulation; not accounting for anticipation will lead to an underestimate of the effect. Second, some land use regulation studies investigate how zoning can cause externalities upon unregulated land values (Zhou et al. (2008), Turner et al. (2014)), such as how curb appeal restrictions improve the value of unregulated but nearby properties. Our study presents an alternative motivation – anticipation about potential rezoning in the future – for why land prices vary systematically in proximity to existing regulation zones.

Evaluating empirically whether anticipated regulatory change affects land values has implications beyond the context of land use regulation. A large literature, like our paper, applies regression discontinuity around administrative boundaries to evaluate public policy. If boundaries in those contexts are known to change over time, then anticipation effects may arise as they do in our study, posing an identification challenge. For example, studies have compared property values across school catchment boundaries to estimate the willingness to pay for school quality (see, for example, Black (1999)).¹ The redrawing of school zone

¹While many regression discontinuity-based studies exploit international or state/provincial boundaries that are likely permanent and thus not subject to the anticipation effects we study, local administrative boundaries are more likely to change over time. For example, municipal boundaries change because of sprawl and annexation of adjacent lands, so comparing outcomes across municipal boundaries may be subject to

boundaries over time is a well-known characteristic of the public school system. To the extent that school boundary changes are anticipated, the housing market will incorporate this information and property values will adjust accordingly.² In this case, only by accounting for anticipated boundary changes will the willingness to pay for school quality be identified.³ More generally, in contexts where the threshold in a regression discontinuity can change over time, anticipation about potential changes to the threshold may in part cause the behavior being studied.⁴

We also contribute to a nascent literature on anticipation effects, which mainly focuses on behavior in anticipation of the implementation of a policy that has already been announced. For example, Coglianese et al. (2016) investigate how gasoline consumption increases prior to the implementation of a rise in the gasoline tax, and how such behavior alters how to use taxes as instrumental variables to estimate the price elasticity of gasoline consumption. Similarly, Lueck and Michael (2003) find that private landowners prematurely harvest their timber to prevent forests from becoming inhabited by endangered species, thereby avoiding costly land use restrictions under the U.S. Endangered Species Act (ESA). In contrast, Malani and Reif (2015) provide a framework for identifying treatment effects in the presence of anticipation of expected future treatment and apply their approach to study how physician labor supply is affected by anticipated reform of tort laws.⁵ Our paper contributes to this literature by studying how to account for and exploit anticipated potential changes of administrative boundaries to evaluate policy.

This paper also contributes to the literature that measures the cost of protecting endangered species. With the exception of Lueck and Michael (2003), mentioned above, this literature focuses on estimating the effects of existing endangered species regulation. For example, Bošković and Nøstbakken (2016), on which this paper builds, find that existing endangered

anticipation effects.

²The possibility of school boundary changes causes some real estate agents to caution clients about such changes. In particular, some advise clients to search for housing strategically within neighborhoods or blocks in order to maximize the chance of being in a desired school zone. For an example, see <http://juliekinnear.com/blogs/school-boundaries-toronto>.

³Though they do not estimate anticipation effects, Ries and Somerville (2010) exploit variation from school catchment rezoning in Vancouver to estimate the effect of school quality on housing prices. Bogart and Cromwell (2000) also examine the effects of school redistricting, and find that redistricting decreased property values in Shaker Heights, Ohio.

⁴In their survey on the use of regression discontinuity in economics, Lee and Lemieux (2010) suggest as much, warning that “optimizing behavior in anticipation of sharp regime change may either accentuate or mute observed effects.”

⁵In contrast to the literature on anticipation effects, Boslett et al. (2016) exploit a lack of anticipation about New York’s moratorium on hydraulic fracturing to measure, using property values, the housing market’s expected net benefits of shale gas development to private landowners in New York.

species regulations impose significant and negative effects on oil lease prices. Greenstone and Gayer (2009) find that ESA critical habitat designations may decrease housing values, while Zabel and Paterson (2006) find that the number of building permits in cities are lower in areas designated as critical habitats for endangered species. Given that endangered species are typically protected using land use regulations and that zoning designations evolve over time, our paper contributes to this literature by providing evidence that anticipated potential changes to endangered species regulations negatively affect the value of land.

The remainder of the paper proceeds as follows. In Section 2, we develop a simple model to guide our empirical strategy. Section 3 describes the data we use. Section 4 describes our empirical specification and reports baseline estimates, which is then followed by an investigation of whether the effect of anticipated regulation varies systematically over space. Section 6 uses the estimation results to calculate total costs of regulation, which is then followed by concluding remarks.

2 A model of anticipated rezoning and land prices

This section describes a one-dimensional spatial model of land prices where land use is regulated and the market anticipates a shift in the geographic boundaries defining where the regulation is applied. The purpose of the model is to shed light on challenges to identifying the effect of regulation on prices when the market anticipates rezoning.

A parcel of land is denoted by x , which lies on the real line. The value of owning a parcel x , exclusive of any costs from regulation, is described by the function $V(x)$, which is unobserved by the researcher. In our application, the owner of x has the right to extract natural resources, so $V(x)$ represents the net present value of profits excluding any costs from land use regulation. Land values are spatially correlated, so that $V(x') \approx V(x)$ for x' sufficiently close to x .

Land use regulation applies to parcels in certain areas or zones. For simplicity, suppose that there is one such zone and the boundary is located at \bar{x} : any parcel x such that $x \geq \bar{x}$ is subject to regulation, while any parcel $x < \bar{x}$ is not subject to regulation. An owner of a regulated parcel incurs a cost, c , which represents the net present value cost of complying with the regulation over the lifetime of owning the parcel.

It is possible that rezoning may occur, implying the boundary may expand or contract. Suppose, for simplicity, that the boundary \bar{x} may expand leftward to \underline{x} .⁶ If there is no

⁶If the zone contracts, so that the boundary shifts to the right, the identification problem is much the

grandfathering with respect to the regulation, then parcels that are unregulated, which lie between \underline{x} and \bar{x} , may be subject to the regulation under rezoning. Commonly-held beliefs about the likelihood of rezoning are described by the probability function $F(x)$, where $F(x) = 0$ for any $x < \underline{x}$ and for any $x \geq \bar{x}$, but $0 \leq F(x) \leq 1$ for any $x \in [\underline{x}, \bar{x})$.

The market for land is competitive, causing the price of a parcel, which is observed by the researcher, to equal the net present value from owning the parcel. In general, the price of a parcel x is given by the function $p(x) = V(x) - F(x)c$. For parcels that are not regulated and for which there is no anticipation that status will change, i.e., $x < \underline{x}$, $F(x) = 0$ so that $p(x) = V(x)$. For parcels that are subject to regulation, $x \geq \bar{x}$, the competitiveness of the market implies that regulatory costs will be capitalized into land prices, so that $p(x) = V(x) - c$ for all $x \geq \bar{x}$. Similarly, leases that are not regulated but may be subject to regulation, $x \in [\underline{x}, \bar{x})$, as a result of rezoning will have the expected cost of anticipated regulation capitalized into their prices: $p(x) = V(x) - F(x)c$. Summarizing the price of land, we have:

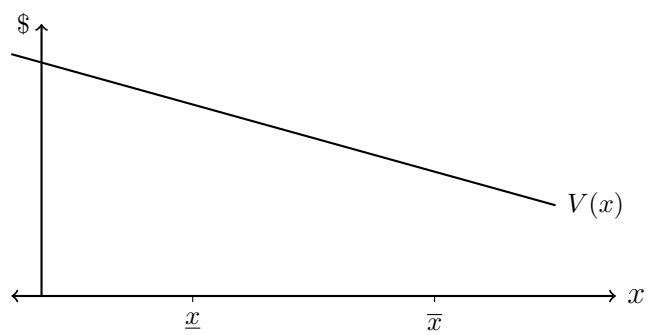
$$p(x) = \begin{cases} V(x) & \text{if } x < \underline{x}; \\ V(x) - F(x)c & \text{if } \underline{x} \leq x < \bar{x}; \\ V(x) - c & \text{if } x \geq \bar{x}. \end{cases} \quad (1)$$

Figure 1 depicts a version of this model. The top panel, Figure 1a, plots land values exclusive of the cost of regulation. For the purposes of illustration, we have depicted $V(x)$ as linearly decreasing along the real line. The next two panels depict the price of land by solid lines. Figure 1b is drawn under the assumption that rezoning will not occur. Figure 1c relaxes this assumption with the additional assumption, for illustrative purposes, that the probability of being subject to rezoning is constant and between 0 and 1, i.e., $F(x) = F$ and $1 > F > 0$ for all $x \in [\underline{x}, \bar{x})$.⁷ In both panels, the dotted line represents the counterfactual price of land if there was no cost of complying with regulation.

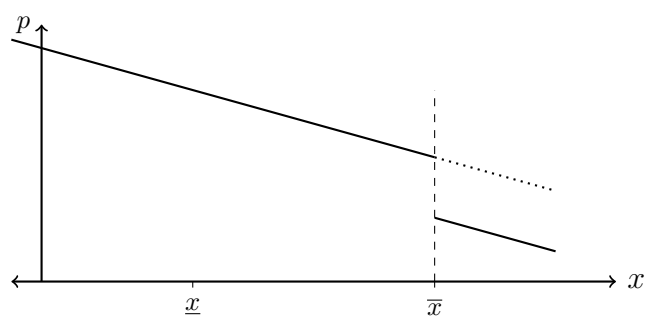
Figure 1b depicts the standard motivation for using regression discontinuity (RD) around geographic borders, with a recent example being Turner et al. (2014). Comparing average land prices on either side of the border will not identify the cost of regulation, c , because parcels are unobservably less valuable in the regulation zone. Due to an assumption that land values are continuous, an RD approach that controls for geographic location by either accounting for proximity to the boundary (thereby proxying for $V(x)$) or by restricting the

same, and unaccounted for anticipation effects will also in this case lead to underestimates of the effect of regulation.

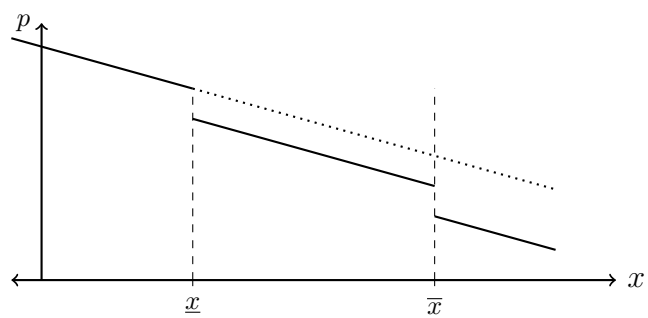
⁷In Section 5, we relax this assumption to test whether parcels nearer to existing regulation boundaries incur greater expected costs than do parcels farther away.



(a) Land values excluding regulatory costs



(b) Land prices with no possibility of rezoning



(c) Land prices with possibility of rezoning

Figure 1: Anticipated rezoning and the price of land

sample to parcels near the boundary identifies the cost parameter.

The scenario in Figure 1c is the focus of this paper. The possibility of rezoning induces the market to capitalize expected costs into the price of parcels for which the status as being unregulated is anticipated to change. While the market may be able to accumulate information about rezoning prospects, such information may be difficult for the researcher to obtain. If rezoning is unobserved by the researcher and the standard RD approach is used, then the cost parameter will not be identified; instead, a difference in prices around the existing boundary will yield $(1 - F(x))c$ for x just to the left of the boundary \bar{x} .

In econometric terms, this identification problem is a misspecification of treatment and control groups: being unregulated, the group of parcels that anticipate regulation are misspecified as part of the control group. The fact that they do incur expected costs from anticipated regulation means their inclusion in the control group will make prices between treatment and control group seem smaller than is the actual cost of existing regulation.

An obvious way to deal with this issue is to account for anticipated boundary changes, so that the omitted variable is controlled for. In doing so, one will have to specify two treatment groups: (1) those parcels facing existing regulation, and (2) those unregulated parcels that may become regulated after rezoning. The existence of the boundary defining the limit of anticipated rezoning, the point \underline{x} in Figure 1c, allows us to define those that anticipate regulation as a separate group from all other parcels. Parcels never subject to regulation, $x < \underline{x}$ in our model, form the control group.

To identify the cost of regulation in practice, one cannot use the subsample of prices for parcels near the regulation boundary. Instead, one must proxy for the unobserved value function $V(x)$ and potentially use the entire sample. Using the example depicted in Figure 1c for illustration, by controlling for the relative location of a parcel by using a linear function of proximity to the boundary \bar{x} , one should be able to control for $V(x)$. In doing so, this effectively separates the land value exclusive of regulatory costs from the price of land. By controlling for which parcels form the two treatment groups – those that are subject to existing regulation and those unregulated parcels anticipating regulation – one can identify the cost of regulation, c . However, unless the probability function $F(x)$ is constant for all parcels in the rezoning area, this approach will not identify $F(x)$. Instead, this approach will identify the average expected cost for unregulated parcels in the rezoning area, $\overline{F(x)}c$, where $\overline{F(x)}$ is the average probability that regulation is anticipated for parcels lying in the rezoning area.

Overall, our analysis shows that anticipated boundary changes will confound identifi-

cation of existing cost of regulation and will yield underestimates of the cost of existing regulation. To deal with this issue requires accounting for anticipated regulation in the empirical analysis. Furthermore, identifying the cost of regulation, as well as the expected cost of anticipated regulation, requires an RD approach that accounts for the geographic location of parcels as controls in price regressions. As we describe in the following sections, we employ this approach in our empirical specifications.

3 Context and data

The setting for our application comes from Alberta, a province in Canada that is rich with underground deposits of hydrocarbons. We make use of auction prices for oil leases and land-use regulations aimed at protecting the environment for which there is reasonable anticipation of rezoning. We describe each of these in turn.

Though mineral rights in Alberta are publicly owned, the provincial government sells development rights to industry. Land leases are sold through first-price, sealed-bid auctions. The government maintains that the auctions for development rights are competitive (Government of Alberta (2009)).⁸ Auctions are announced publicly ahead of the auction date, and the winner and winning bid for each auctioned parcel are made public immediately. Leases can be held indefinitely so long as they continue producing.

Predominant among Albertan hydrocarbons are oil sands, a mix of bitumen, water, and sand from which synthetic crude oil can be produced. As of 2013, the prevalence of oil sands puts Alberta as having the third-highest reserves of oil in the world (ERCB (2013)). Oil sands production requires large start-up and operating costs relative to conventional sources, which make oil produced from sands the marginal barrel for world oil supply (International Energy Agency (2015)). The nearly 4,000 oil sands leases sold between 2003 and 2012, the time period for our sample, generated about \$4 billion in revenue (all dollar figures in this paper are in 2012 Canadian dollars). The average price for these leases is about \$800 per hectare; Figure 2 plots their geographic price distribution.

The development of oil sands imposes significant negative effects on the environment. Though known for its high greenhouse gas emissions, oil sands production creates immense land disturbances. Among other impacts, this has caused and continues to cause severe population declines in woodland caribou, a species designated as endangered in Alberta (Boutin

⁸Watkins (1975) and Watkins and Kirby (1981) found that Alberta’s oil and gas auctions were competitive; the industry has only grown since the time of these studies.

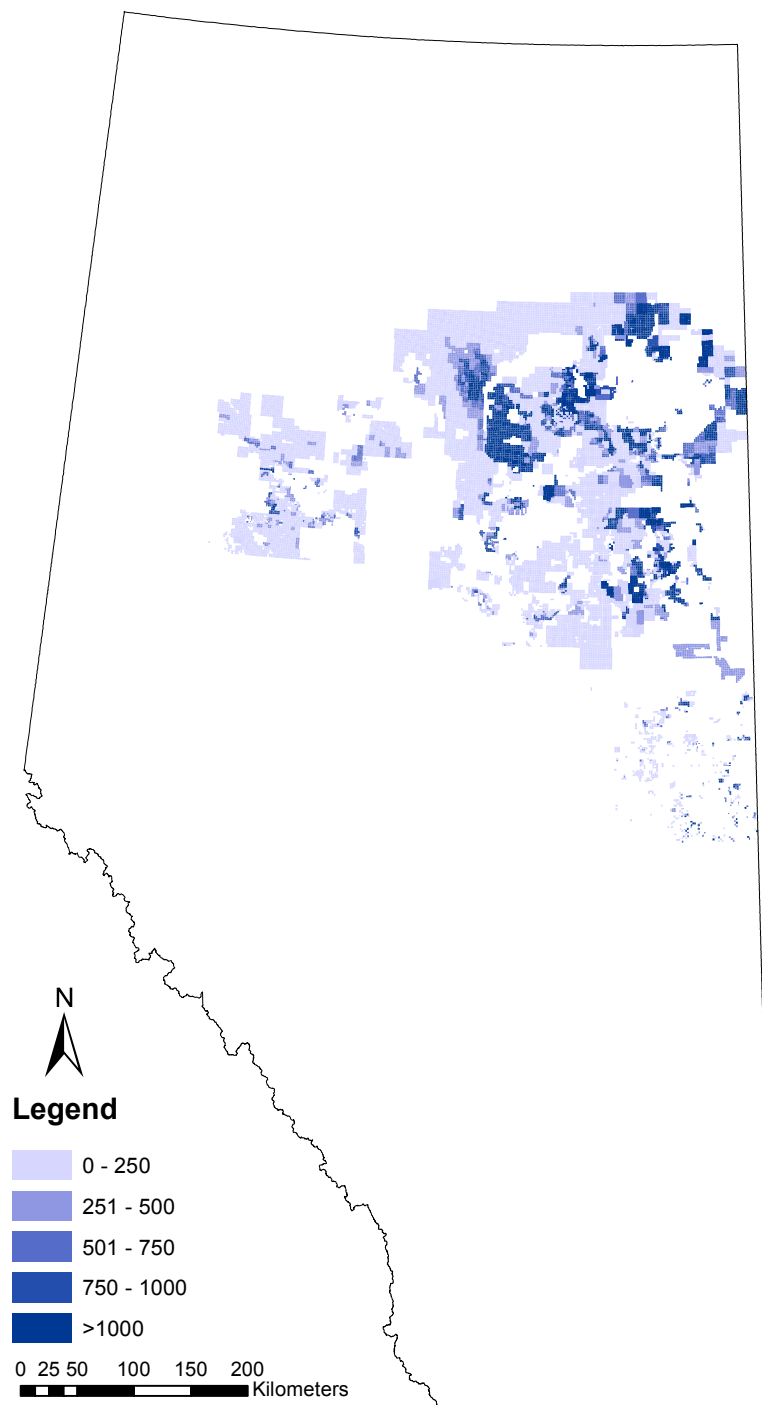


Figure 2: Price per hectare, in 2012 Canadian dollars, of oil sands leases auctioned during 2003–2012

et al. (2012)). To protect the endangered caribou from continued risk, the provincial government has imposed land-use regulations in caribou protection zones that restrict development. The 31 geographically distinct zones, depicted in the darker shade in Figure 3, are drawn based on critical caribou habitat areas for herds. An owner of a lease located in one of the zones must develop a strategic plan to mitigate adverse effects on caribou habitat and migration, and the regulator must approve the plan. Operationally, the zones impose constraints on production, implying lease holders incur costs from complying with the regulation that would not be incurred outside the zones.⁹

While a protection zone is based narrowly on a herd’s critical habitat, the associated caribou range defines the broader area in which the herd migrates and lives.¹⁰ As can be seen from Figure 3, the caribou zones, the darker shade, are geographic subsets of caribou ranges, depicted in the lighter shade. The reason this distinction is important is because the zones, since their inception in 1991, have been expanded incrementally in the direction of range boundaries and have never been contracted. The boundaries of caribou ranges, describing the outer limits of a herd’s habitat and migratory areas, provide a natural endpoint for zone expansion. Indeed, in 2013 the government expanded the zones a final time: protection zones are now defined by range boundaries everywhere.

Even though a lease may lie in a range, as long as it lies outside zone boundaries it is not subject to the regulation. A firm bidding on such a lease will know that, because of the history of caribou zone expansion, its desired lease faces a positive probability of being rezoned as part of an expanded protection zone. The firm will incorporate its belief about this likelihood and the net present value of complying with the regulation into its bid for the lease. As a result, this setting is an appropriate context for testing the model’s predictions in Section 2 by producing the three categories of land: (1) the protection zones, where leases are currently subject to regulation; (2) the range areas outside protection zones, where lease holders can reasonably expect to be subject to regulation following rezoning, and; (3) the area outside both range and zone boundaries, which will never be regulated.

We follow this categorization by overlaying the auctioned leases on the map of regulation zones and ranges to determine whether a given lease falls within a particular zone or outside the zone but within the associated range. We categorize any lease falling outside any zone

⁹Examples of such activities include reducing the clear-cutting of forests, creating roads and pipelines that circumvent caribou routes and habitats, limiting seismic disturbances from drilling, and restoring seismic lines and cleared areas to original conditions as soon as possible.

¹⁰The ranges were mapped from surveys in the late 1970s and early 1980s (Alberta Woodland Caribou Recovery Team (2005)). And, unlike boreal caribou, woodland caribou do not migrate long distances.

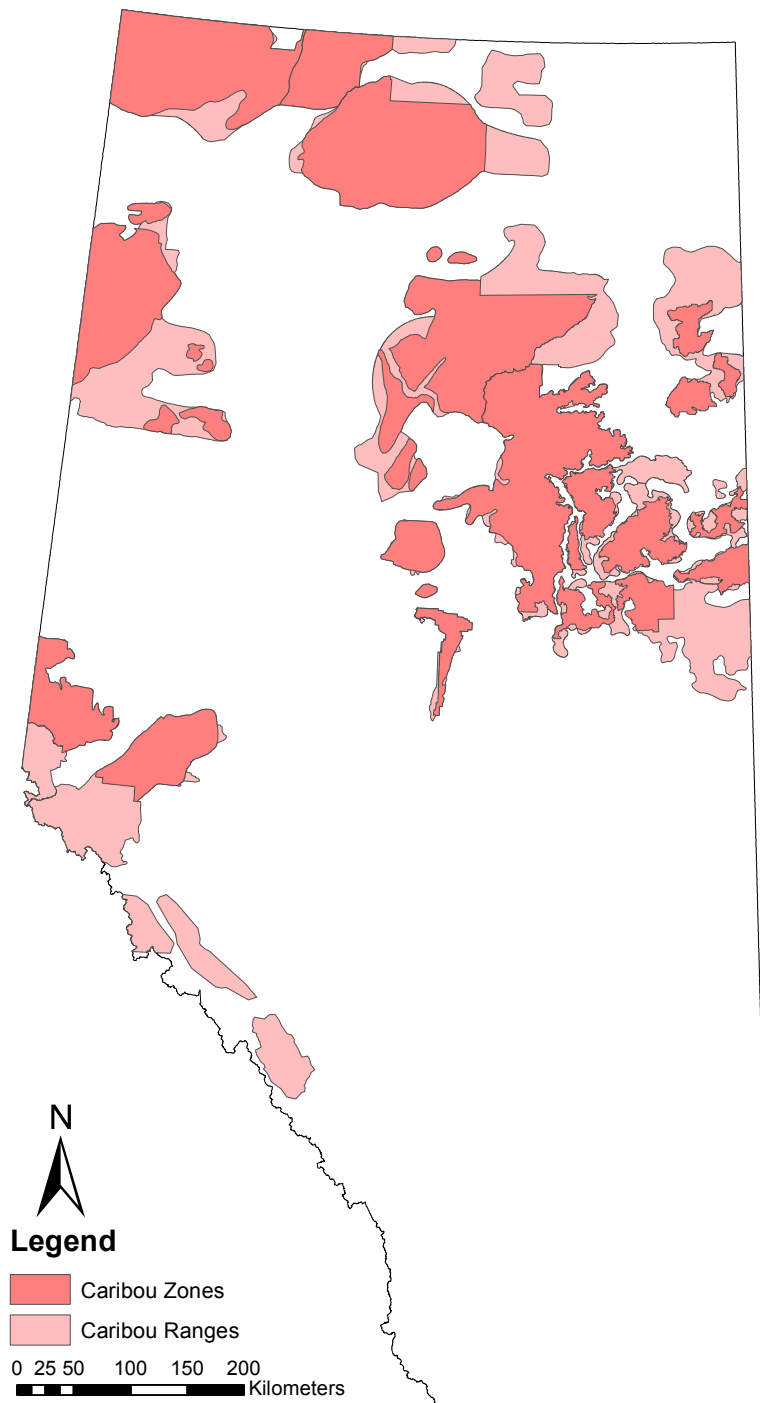


Figure 3: Caribou zones and ranges, 2003–2012

and any range as never being regulated, either now or in the future. We exclude any leases that fall into multiple zones.

For each lease, we determine its geographic centroid, in degrees latitude and longitude, to calculate distances to zone boundaries. We exclude any leases more than 200 kilometers outside a zone or range and any leases overlapping a boundary. Also, for estimation purposes, we exclude any lease whose owner is observed only once in our sample. We also focus on leases sold between 2003 and 2012, since the boundaries were shifted in 2002 and again, as mentioned above, in 2013.¹¹ Altogether, these criteria narrow our sample to 2,918 leases that lie inside or near 17 protection zones and 13 ranges.

Table 1 reports sample means and standard deviations for leases for the full sample in the first column, and by regulation status in columns 2 to 4. Also reported in the final three columns are difference in means for each subsample of regulation status and the associated standard error in parentheses. The number of regulated leases – those located inside protection zones – equals 1,326 in our sample, which is roughly equal to the number of leases that are and never will be regulated – those located outside zone and range boundaries – of which there are 1,272. There are 320 leases for which rezoning is anticipated, which constitutes approximately 10% of the sample.

The first row reports statistics for the auction price per hectare of individual leases. The price measure is higher on average for leases within protection zones than everywhere else. The difference between regulated and the unregulated range prices per hectare is statistically significant, while the other differences are not. The distribution of price per hectare for our sample has a long right tail; this fact motivates us to use the natural logarithm of price per hectare for our empirical analysis in the next section.

The second row reports summary statistics for the count of firms with an ownership stake in a given lease. On average most leases are owned by a single firm.¹² The total number of firms in the sample is 95, and the number of firms in a given group is reported in the bottom row of the table.¹³ In the full sample, 68 of the 95 firms own leases inside zone boundaries,

¹¹We obtained the location of the zones and ranges from shapefiles available on the Government of Alberta’s website. Prior maps are not available in digital form and repeated requests to the government for precise maps have proved unsuccessful. The expansion in 2013 was followed by an intervention by the federal government, under the auspices of the federal *Species at Risk Act (SARA)*, into saving the caribou. Under *SARA*, new rules are due in 2016; the uncertainty of what rules will be implemented caused us to cut off the sample at 2012.

¹²If a lease is owned by more than one firm, we will refer to the owner that interacts with the regulator as the main owner.

¹³As we mentioned above, we excluded from our sample any leases for which we observe the owner only once. This criterion removes 26 unique firms from our sample and does not affect the distribution of firms

Table 1: Summary statistics for oil sands leases by regulation status

	Full sample	Regulation status:			Difference in sample means:		
		Zone	Range	Neither	Zone - Range	Zone - Neither	Range - Neither
Price/hectare (2012 CDN \$)	768.92 (1885.85)	801.22 (1753.94)	664.92 (1505.39)	761.41 (2093.44)	136.30 (106.41)	39.81 (75.65)	-96.50 (124.41)
Number of joint owners	1.10 (0.36)	1.10 (0.35)	1.16 (0.51)	1.08 (0.33)	-0.06** (0.02)	0.02 (0.01)	0.08*** (0.02)
Special access area	0.02 (0.13)			0.04 (0.19)			
Key wildlife area	0.13 (0.34)	0.05 (0.22)	0.21 (0.41)	0.211 (0.407)	-0.10*** (0.02)	-0.16*** (0.01)	-0.05** (0.02)
Depth (meters)	828.76 (110.74)	803.63 (61.96)	824.36 (101.74)	856.08 (141.89)	-20.72*** (4.45)	-52.44*** (4.27)	-31.72*** (8.43)
Number of firms	94	66	49	88			
Observations	2918	1326	320	1272	1646	2598	1592

Notes: The first four columns report means and standard deviations, in parentheses, of variables in the full sample, for leases within zones, for leases within ranges but outside zones, and for leases outside of both zones and ranges, respectively. The next three columns report sample mean differences and standard errors for: (1) leases in zones and leases in ranges; (2) leases lying in zones and leases outside zones and ranges, and; (3) leases in ranges and leases outside zones and ranges. *, **, and *** denote difference in means different from zero at the 10%, 5%, and 1% significance levels. The final two rows report the number of unique firms in each sample and the total observations in each sample, respectively.

49 firms own leases in unregulated range areas, while 88 of the firms own leases outside of the zones.

Rows three and four report summary statistics for binary variables indicating whether the lease falls into areas where other forms of regulation are applied, known as key wildlife and special access areas, which impose different land-use regulations to protect biodiversity.¹⁴ The final row reports summary statistics for the depth, in meters, of the core sample for that lease.¹⁵ This variable provides information on the cost of extracting oil sands for that lease, since deeper deposits require greater effort and more cost to produce to the surface. This is consistent with the correlation with average prices: leases outside zones are deeper on average and command lower prices per hectare than leases in zones.

4 Baseline empirical specification and results

In this section, we describe the empirical specification for estimating the effects of the current caribou protection regulation and the anticipated rezoning of unregulated areas on auction prices for oil sands leases. In the following subsections, we discuss identification, report our baseline estimation and sensitivity results, and report and provide a discussion of results from failing to account for anticipated rezoning.

To estimate the effect of existing and anticipated regulation on auction prices, we follow the analysis in Section 2 by specifying a regression that exploits the discontinuity in regulation and anticipated regulation at zone and range boundaries, respectively. The model makes clear that accounting for geographic location is critical for identification. To do so in two-dimensional space, we follow Dell (2010) by employing a multidimensional RD that accounts for geographic location using latitude and longitude as well as the identity of the nearest zone. In particular, for lease i near zone z , purchased by firm j in year t , we estimate:

$$p_{izjt} = \alpha + \beta_1 \text{zone}_i + \beta_2 \text{range}_i + W_i \Gamma + f(x_i, y_i) + \lambda_z + \lambda_{jt} + \varepsilon_{izjt}. \quad (2)$$

The dependent variable, p_{izjt} , is the logarithm of the auction price per hectare for lease i . The variable zone_i describes whether lease i is subject to the land use regulation: the variable is equal to 1 if lease i is located inside any zone, implying that the lease is regulated, and is

across the areas of different regulation status.

¹⁴See Government of Alberta (2013) for more details on these regulations. Conversations with some of the producers revealed that complying with the caribou land-use regulation is of much greater concern.

¹⁵Core samples are made publicly available by the Government of Alberta; see <http://www.ags.gov.ab.ca/services/mcrf/>.

equal to 0 otherwise, in which case it is unregulated. The variable range_i is equal to 1 if lease i lies inside a caribou range but outside any caribou protection zone, and is otherwise equal to 0. This indicator variable defines the group of leases that are currently unregulated but can reasonably anticipate, given the history of these regulations, to be subject to regulation in the future following an expansion of the zones. Should both zone_i and range_i both equal 0, then lease i lies outside of any zone or range and is not and never will be regulated. Though leases for which $\text{range}_i = 1$ are unregulated, for convenience we will refer to them as “range leases” to distinguish them from unregulated leases located outside ranges, which we will refer to as “unregulated leases.”

The multidimensional RD is based on the variables x_i and y_i , describing the latitude and longitude in degrees of lease i , the function f , and the parameters λ_z . The unknown function f controls for smooth functions of geographic location that affect the profitability of land and, thus, the observed winning bid for lease i , such as the underlying resource stock and distances to hubs where producers obtain their inputs for production. The parameters λ_z are fixed effects that account for the zone to which lease i is near, and control for differences in lease prices across zone regions.¹⁶

A practical issue is how to estimate equation (2) given the function f is unknown. Here we again follow Dell (2010) and specify f as a series of polynomials in latitude and longitude. While there is no rule on which degree polynomial to employ, though too high a degree may result in overfitting, we use a quadratic polynomial for our baseline specification.¹⁷ To test the sensitivity of our results to this specification, we will report results from re-estimating equation (2) using polynomials of lower and higher degree.

The vector W_i contains information specific to the lease, such as the number of joint owners, all of which are listed in Table 1. The parameters λ_{jt} are fixed effects that account for time-varying firm-specific differences in winning auctions bids that account for, among other factors, differences in production technologies and the timing of when firms obtain leases.

The parameters β_1 and β_2 are the parameters of interest. The parameter β_1 captures the price effect for a lease that is regulated because it is located inside a protection zone. If the regulation imposes costs over the lifetime of owning the lease and such costs are capitalized

¹⁶Recall that each zone is a subset of a range and that multiple zones might fall into a range. As a result, employing a range fixed effect is less flexible than employing zone fixed effects and range fixed effects are redundant given we use zone fixed effects.

¹⁷If latitude and longitude are denoted by x and y , then the quadratic in latitude and longitude is $x^2 + y^2 + xy + x + y$.

into the winning bid for the lease, then we expect that the regulation will impose a negative effect on price, so that $\beta_1 < 0$.

The parameter β_2 captures the average price effect of owning a lease that is located inside a range but outside any protection zone.¹⁸ The latter implies the lease is not regulated, while the former implies there is reasonable anticipation it will be subject to regulation in the future following rezoning of the protection zones. If the anticipated regulation imposes expected costs and those costs are capitalized into lease prices, then we expect that $\beta_2 < 0$. If bidders believe that the probability that such a lease will be rezoned is close to 1 and that rezoning will occur sufficiently soon, then the expected cost of anticipated regulation will approximately be equal to the cost of existing regulation, in which case we expect that $\beta_2 = \beta_1$. Otherwise, the expected costs should be lower than the cost of regulation incurred by a lease inside a zone, so that $\beta_2 > \beta_1$.

4.1 Identification

The identification strategy premised on equation (2) attributes, after controlling for geographic location, firm-specific time-varying heterogeneity, and lease-specific differences, any difference in prices between leases inside zones and unregulated leases to be caused by the regulation. Similarly, after incorporating our full set of controls, we attribute any price difference between range leases and unregulated leases to be caused by anticipation of potential future regulation. Several issues affect the identification and interpretation of the effect of existing and anticipated regulation, such as: (1) the determination of zone and range boundaries; (2) selection of leases by firms; (3) externalities caused by the regulation, and; (4) competitiveness of auctions. We discuss each of these in turn.

An implicit identification assumption we are making is that the zone and range boundaries are determined independently of the underlying oil sands resource. If that were not the case, then we may be erroneously attributing any difference in prices as being caused by existing or anticipated regulation, whereas the difference may be due in whole or in part to the unobservable resource quality associated with the leases that discontinuously changes at zone or range boundaries. However, we have no reason to believe the boundaries were determined by the location of the resource. The range boundaries, describing the outer limits of a caribou herd's habitat and migration area, were drawn by ecologists decades prior to the initial implementation of the land use regulations (Dzus (2001)). The zones describe critical

¹⁸In Section 5, we investigate whether the effect of anticipated rezoning varies with the likelihood of being rezoned, which we measure by proximity to the zone boundary.

habitat for caribou, which is determined mainly by the location of old-growth boreal forests on which caribou rely for survival (Environment Canada (2012)).¹⁹ Such boreal forests cover not only most of the northern half of Alberta, but also the northern area of most Canadian provinces, and woodland caribou exist in most of those regions. It is a coincidence that oil sands deposits lie below caribou ranges and critical habitats in Alberta.

The second potential identification issue is whether leases are chosen based on whether they are located in zones or ranges; the ability to manipulate the regulation treatment by selecting a lease based on location would invalidate a typical RD application. However, in our application, regardless of whether the lease is inside a zone or a range or is unregulated, in equilibrium bidders will pay the value of owning that land. As such, a firm that chooses to purchase a lease in a regulated area can fully compensate itself for the cost of complying with the regulation by decreasing its bid in an amount equal to the cost and still win the lease. This behavior does not pose an identification problem, but instead we rely on this behavior to identify the effect of existing and anticipated regulation on land prices.²⁰

A similar issue is whether some leases that would have been purchased in the absence of regulation are left unsold because the regulation causes them to be unprofitable. We are unable to deal with this selection issue, in which case we will be identifying the effect of existing and anticipated regulation on leases for which the net present value of owning the lease, including the cost of regulation, is non-negative.²¹

The third identification issue is externalities arising from the regulation. The presence of such externalities motivates part of the analysis in Turner et al. (2014), who study how land use regulation for residential housing, such as restrictions that homeowners maintain curb appeal, can impose externalities on prices for housing across a nearby administrative boundary. If such externalities exist, then they may cause a similar effect on lease prices as does anticipation about potential rezoning.

We do not suspect externalities to pose an identification problem for two reasons. The first reason is based on context: externalities such as those present in the zoning of residential housing are unlikely to arise in our context. Aside from the spatial correlation in resource

¹⁹Oil sands extraction projects do not necessarily involve forest removal and land clearing: *in situ* projects, which are the more prevalent form of oil sands operations, create linear land disturbances by creating access roads but otherwise keep much of the forest intact.

²⁰It may be the case that certain firms are better able – in the least-cost sense – to comply with regulations and so are more likely to obtain leases in the zones. However, assuming auctions are competitive, we are able to control for producer identity, and thus our identification strategy is robust to this issue.

²¹We discussed this issue with several members of the industry, who claimed that the costs of complying with the caribou protection regulation were not large enough to dissuade the purchase of any lease their firm was considering.

stocks, how one oil sands firm bids for a lease should not be affected by the fact that nearby leases are regulated. The second reason is due to our identification strategy. If externalities can arise in our context, then because both unregulated leases and range leases are located near zones, both will equally be subject to externalities. Our strategy for identifying the effect of anticipated rezoning compares prices for range and unregulated leases, in which case the effect of externalities should be differenced out. As a result, we do not suspect that this should confound our estimates of the effect of existing anticipated regulation. Nonetheless, in the following subsection, we perform robustness checks to determine externalities may be driving our baseline estimates.

The final identification issue we consider is the competitiveness of the auctions. Although we observe the winning auction bid, we do not observe, and neither does anyone but the regulator, any non-winning bids or the number of bidders. The government claims that the auctions are competitive (Government of Alberta (2009)) and there is no evidence suggesting otherwise. Watkins (1975) and Watkins and Kirby (1981) found that Alberta’s oil and gas lease auctions in the 1970s were competitive, and the oil and gas industry in Alberta has only grown since then.

4.2 Baseline estimation results

The results from estimating equation (2) using progressively richer specifications are reported in Table 2. Column (1) reports coefficient estimates from a regression that controls only for whether a lease is located inside a zone or a range. The coefficient for the zone variable, 0.0289, indicates that a lease inside a zone is associated with a price per hectare that is 2.9% larger on average than the price per hectare for an unregulated lease. The coefficient estimate for the range variable, -0.0199 , indicates that a lease in a range is associated with a price per hectare that is nearly 2% lower on average than an unregulated lease. These estimates reflect the pattern of unconditional mean prices per hectare reported in Table 1.

Columns (2) and (3) add the polynomial in latitude and longitude and the zone fixed effects, respectively, that form the multidimensional RD. The addition of the latitude-longitude polynomial in column (2) decreases the value of both zone and range coefficients so much that both estimates are negative and statistically significant. The estimates reported in column (3) include the zone fixed effects, which control for the identity of the nearest zone to a given lease; their inclusion reduces the magnitude of both coefficient estimates relative to column (2), although much less so for leases lying in zones.

Column (4) adds the firm-year fixed effects, which are important not only in controlling

for time trends in this volatile industry but also in how each firm responds idiosyncratically to those trends. These fixed effects control for a significant amount of variation in the dependent variable, which can be seen from the substantial increase in the R^2 value. The coefficient estimate for the zone variable, -0.3179 , is statistically significant at the 1% level, while the coefficient estimate for the range variable, -0.1535 , is insignificant at conventional levels.

Table 2: Estimates of the effects of existing and anticipated regulation on lease prices

	(1)	(2)	(3)	(4)	(5)
Zone	0.0289 (0.1743)	-0.3326** (0.1362)	-0.2254* (0.1144)	-0.3179*** (0.0976)	-0.3536*** (0.0878)
Range	-0.0199 (0.2085)	-0.5339*** (0.1737)	-0.2578 (0.1901)	-0.1535 (0.1173)	-0.1721 (0.1174)
Number of joint owners					-0.1227 (0.2685)
Key wildlife area					-0.1756*** (0.0391)
Special access area					0.3183 (0.4047)
Log(depth)					-0.2284 (2.6386)
Quad. poly. in lat/lon	N	Y	Y	Y	Y
Zone fixed effects	N	N	Y	Y	Y
Owner-Year fixed effects	N	N	N	Y	Y
R^2	0.0001	0.1002	0.1549	0.5534	0.5543
Observations	2918	2918	2918	2918	2918

Notes: The dependant variable is the logarithm of price per hectare at the lease level. Robust standard errors, adjusted for clustering by caribou zone, in parentheses. *, **, and *** denote estimates different from zero at the 10%, 5%, and 1% significance levels.

The estimates in column (5) are from the full specification, which adds the lease-specific controls listed in Table 1. These controls do not change the estimates much numerically relative to the column (4) estimates. The coefficient estimate for the number of joint owners, listed in the fourth row, indicates that the price per hectare for a lease decreases with the size of group ownership over a lease; this variable likely proxies for risk, as the practice in this industry is to share risk by entering into joint ownership. The coefficient estimate for the key wildlife area is negative and statistically significant, indicating that the price for a lease subject to the key wildlife land use regulations is lower than the price for a lease that is not in that area. The opposite is true for special access areas, though these are imprecisely

estimated given few leases in our sample lie in these areas and the ones that do are all unregulated leases. The coefficient estimate for the logarithm of core depth is statistically insignificant, but being of negative sign is consistent with the notion that deeper deposits are costlier to extract.

Because it uses the full set of controls, the specification in column (5) is our preferred one, so the estimates of interest in that column deserve a closer look. The coefficient estimate for the zone variable, -0.3536 , is statistically significant at the 1% level. The estimate indicates that the price per hectare for a lease located inside a zone and thus subject to the endangered caribou regulations is 30% lower on average than the price per hectare for an unregulated lease. At the mean and median prices per hectare in our sample of \$803 and \$135, this estimated effect amounts to a decrease in price of a lease by \$238 and \$41 per hectare. We take the estimated negative effects on the price of zone leases as evidence that the land use regulations impose costs on producers.

The coefficient estimate reported in Table 2 for the range variable, -0.1721 , is not significant at conventional levels, but has a p -value of 0.162 even though our sample contains only 320 ranges leases. The estimate implies that the price per hectare for a lease in a range, even though it is not regulated, is about 16% lower on average than the price per hectare for an unregulated lease. At the mean and median price per hectare, the estimate implies a price discount of approximately \$128 and \$22 per hectare, respectively. The negative price effect for range leases is consistent with the prediction that the expected costs of anticipated future regulation are capitalized into auction prices for those leases.

As we described above, we arbitrarily chose a quadratic polynomial to approximate the unknown function f in equation (2). To evaluate how sensitive our estimates are to the choice of latitude-longitude polynomial, we re-estimate equation (2) using polynomials of lower and higher degree. Table 3 reports coefficient estimates from using a linear polynomial, in Panel A, and a cubic polynomial, in Panel B.²² The columns correspond to the column in Table 2 with the same set of controls. As a result, a given coefficient estimate can be compared to the coefficient estimate in Table 2 with the same set of controls, as well as the same-column estimate but for different polynomial specifications in Table 3.

The estimates in Panel A of Table 3 for the zone variable are lower in magnitude than the corresponding estimate in Table 2. In column (4) of Table 3, which uses the full set of controls, the estimate of -0.2885 implies that the price per hectare for a lease inside a zone

²²We were unable to use polynomials of degree 4 or higher because, given the rich set of fixed effects that we employ and the geographic concentration of leases by firms and over time, using polynomials of such a high degree causes collinearity problems.

is 25% lower than the price per hectare for an unregulated lease. This implied price discount is smaller than the 30% price discount implied by the estimate in column (5) of Table 2, likely due to underfitting.²³ The coefficient estimate for the range variable in column (4), -0.1735 , is similar numerically to the estimate from the preferred specification in column (5) of Table 2.

Table 3: Sensitivity of estimates using different latitude-longitude polynomials

	(1)	(2)	(3)	(4)
Panel A: Linear polynomial				
Zone	-0.2697 (0.1865)	-0.1296 (0.1058)	-0.2706** (0.1048)	-0.2885** (0.0999)
Range	-0.5189** (0.1831)	-0.3230* (0.1631)	-0.1618 (0.1152)	-0.1735 (0.1181)
R^2	0.0885	0.1388	0.5519	0.5525
Panel B: Cubic polynomial				
Zone	-0.2709* (0.1322)	-0.3272** (0.1126)	-0.3258*** (0.1047)	-0.3563*** (0.1046)
Range	-0.3703** (0.1603)	-0.2635* (0.1491)	-0.1227 (0.1331)	-0.1444 (0.1381)
R^2	0.1319	0.1850	0.5534	0.5543
Zone fixed effects	N	Y	Y	Y
Owner-Year fixed effects	N	N	Y	Y
Lease controls	N	N	N	Y
Observations	2918	2918	2918	2918

Notes: The dependant variable is the logarithm of price per hectare at the lease level. Robust standard errors, adjusted for clustering by caribou zone, in parentheses. *, **, and *** denote estimates different from zero at the 10%, 5%, and 1% significance levels.

The coefficient estimates for the zone variable when using a cubic latitude-longitude polynomial, reported in Panel B, are more similar in value to our baseline estimates relative to estimates from using a linear polynomial. In particular, once controlling for the main components of the multidimensional RD – the latitude-longitude polynomial and the zone fixed effects, in column (2) – the zone coefficient estimate remains stable as we add more controls. The coefficient estimate in column (4), -0.3563 , is virtually the same as the preferred estimate in column (5) of Table 2. Similarly, the coefficient estimate for the range variable is quite close in value to the estimate in column (5) of Table 2.

²³We found that our baseline estimates can be closely replicated by controlling for latitude and longitude individually to a power greater than 2, as well as any interaction of any power of latitude and longitude.

Overall, we take the estimates in Table 3 as evidence that our baseline estimates are robust to the choice of polynomial in latitude and longitude, so long as we employ a polynomial of appropriately high degree.

The second robustness check we perform is for the presence of externalities caused by the regulation that may affect nearby leases outside zone boundaries. As we discussed in Section 4.1, our strategy to identify the effect of anticipated regulation will, if unregulated leases and range leases are equally affected by externalities (because of their proximity to zone boundaries), difference out any effect of externalities on lease prices. However, in our sample, unregulated leases are on average farther from zone boundaries than range leases. Because externalities caused by the regulation likely dissipates with distance from zone boundaries, unregulated lease prices are on average less affected by externalities than range lease prices. As a result, our comparison of range and unregulated lease prices may not fully difference out the effect of externalities, since on average range and unregulated leases are not equally affected by externalities. Though we do account for geographic location in our estimation specification, likely controlling for this issue, a simple way to determine whether this is an issue in practice is to restrict our sample so that range and unregulated leases are of similar proximity to zone boundaries and then re-estimate equation (2). We do this by restricting range and unregulated leases to be no farther from zone boundaries than 75, 50, and 25 kilometers, and report estimation results in Table 4. Each panel reports a different set of estimation results using the specified distance cutoff, and the specification in each column is identical across panels and is identical to the same-numbered column in Table 2, making the estimates directly comparable.

The estimates in each panel are fairly similar to the baseline estimates reported in Table 2. In particular, consider the restricted sample using range and unregulated leases that are no farther than 75 kilometers from the nearest zone boundary, in Panel A. The coefficient estimate in column (5), -0.1762 , implies a price per hectare discount for a range lease is 16% on average; this estimated price effects is virtually the same as the price effect implied by the coefficient estimate, -0.1721 , in column (5) of Table 2. The estimates in Panels B and C, restricting the sample to include range and unregulated leases no farther than 50 and 25 kilometers from the nearest zone boundary, are virtually the same. Based on those coefficient estimates for range leases, the price per hectare discount is 15% on average. Since the coefficient estimates for range leases reported in Table 4 are virtually identical to the baseline estimates in Table 2, we take this as suggestive evidence that our baseline estimates for the effect of anticipated regulation are not due to externalities caused by the regulation.

Table 4: Sensitivity of estimates based on proximity of unregulated and range leases to nearest zone boundary

	(1)	(2)	(3)	(4)	(5)
Panel A: Within 75 km of nearest zone boundary ($n = 2446$)					
Zone	-0.0691 (0.2022)	-0.3645** (0.1445)	-0.3662*** (0.1232)	-0.3939*** (0.1072)	-0.4056*** (0.1145)
Range	-0.1178 (0.2258)	-0.5032*** (0.1370)	-0.3223** (0.1485)	-0.1883 (0.1265)	-0.1762 (0.1244)
R^2	0.0004	0.1030	0.1679	0.5680	0.5710
Panel B: Within 50 km of nearest zone boundary ($n = 2274$)					
Zone	-0.2903 (0.2018)	-0.4162** (0.1616)	-0.3977*** (0.1212)	-0.4088*** (0.1202)	-0.4056*** (0.1252)
Range	-0.3464* (0.1921)	-0.5305*** (0.1427)	-0.3513** (0.1212)	-0.1832 (0.1293)	-0.1688 (0.1300)
R^2	0.0050	0.1000	0.1624	0.5756	0.5796
Panel C: Within 25 km of nearest zone boundary ($n = 2099$)					
Zone	-0.4267** (0.1665)	-0.4376** (0.1671)	-0.3981*** (0.1074)	-0.4130*** (0.1196)	-0.3985*** (0.1190)
Range	-0.5287*** (0.1509)	-0.5384*** (0.1378)	-0.2941** (0.1102)	-0.1878 (0.1250)	-0.1653 (0.1224)
R^2	0.0096	0.0916	0.1591	0.5724	0.5775
Quad. poly. in lat/lon	N	Y	Y	Y	Y
Zone fixed effects	N	N	Y	Y	Y
Owner-Year fixed effects	N	N	N	Y	Y
Lease controls	N	N	N	N	Y

Notes: The dependant variable is the logarithm of price per hectare at the lease level. Robust standard errors, adjusted for clustering by caribou zone, in parentheses. *, **, and *** denote estimates different from zero at the 10%, 5%, and 1% significance levels.

4.3 On the bias from not accounting for anticipated rezoning

Rezoning is a common feature of geographically administered policies. While our context allows us to account for which leases located outside regulation zones may be subject to regulation in the future due to rezoning, it may be difficult to obtain information on whether rezoning is likely or precisely where rezoning may take place. As predicted by the model in Section 2, not accounting for which land parcels may be rezoned leads to an underestimate of the effect of regulation on lease prices. To test this prediction, we estimated a version of equation (2) that omits the variable indicating which leases are in range areas. If the prediction is true, then the coefficient estimate for a regulated lease from a specification that omits the variable range_i will be smaller in magnitude than the estimate from our baseline specification.

We report the estimation results from a specification that omits the range variable in Table 5. The specification in each column is identical to the column of the same number in Table 2, so the estimates are directly comparable.

Across the specifications in each column, the coefficient estimates describing the effect on the price for a lease located inside a protection zone in Table 5 are consistently lower in magnitude than the corresponding estimates in Table 2. This finding supports the prediction that failing to account for anticipation about rezoning leads to underestimates of the effect of regulation on lease prices. Focusing on column (5) of Table 5, the coefficient estimate of -0.2879 implies a negative price per hectare effect of 25% for leases inside zones, which is 5 percentage points smaller in magnitude than when we control for which leases located in ranges. To put this into perspective, the price per hectare effect based on our baseline estimates, from the mean price per hectare, is \$238; this amount is underestimated by \$38 per hectare when not accounting for which leases may plausibly be rezoned in the future.

We can further explore the size of the bias due to not accounting for anticipated rezoning by exploiting the fact that for some zones the boundary coincides or is close to the associated range boundary, while for other zones the boundaries are distant from one another. While still not accounting for anticipated rezoning by omitting the range variable, we do this by estimating a specification of the following form:

$$p_{izjt} = \nu + \delta_1 \text{zone}_i + \delta_2 \text{zone}_i \times \text{size}_z + W_i \Pi + f(x_i, y_i) + \mu_z + \mu_{jt} + \epsilon_{izjt}, \quad (3)$$

where size_z is a measure of the size of a range associated with zone z . This estimating equation is similar to a version of equation (2) that omits range_i , but now the (biased) effect

Table 5: Estimates reflecting the bias from not accounting for anticipation about rezoning

	(1)	(2)	(3)	(4)	(5)
Zone	0.0329 (0.1489)	-0.1465 (0.1126)	-0.1346 (0.0886)	-0.2604** (0.1046)	-0.2879*** (0.0905)
Quad. poly. in lat/lon	N	Y	Y	Y	Y
Zone fixed effects	N	N	Y	Y	Y
Owner-Year fixed effects	N	N	N	Y	Y
Lease controls	N	N	N	N	Y
R^2	0.0001	0.0942	0.1536	0.5531	0.5539
Observations	2918	2918	2918	2918	2918

Notes: The dependant variable is the logarithm of price per hectare at the lease level. Robust standard errors, adjusted for clustering by caribou zone, in parentheses. *, **, and *** denote estimates different from zero at the 10%, 5%, and 1% significance levels.

of lying in a zone is equal to $\delta_1 + \delta_2 \text{size}_z$. Our expectation is that larger range areas which may be rezoned in the future (and for which anticipated rezoning is unaccounted for), leads to greater bias in estimates of the price effect of being located inside a zone. The inclusion of the interaction term allows us to test this prediction, with the expectation that larger ranges are associated with greater bias, implying $\delta_2 > 0$.

We use two different measures for size_z . The first is the total count of leases sold in a range area near a given zone over the course of the entire sample. If there are a greater number of leases near a zone for which rezoning is anticipated and the expected costs of rezoning are capitalized into their lease prices, we expect the omission of their anticipated regulation status to dilute the effect of regulation than if there are no leases near a zone for which regulation is anticipated. The second measure is the distance of the lease in a range that is farthest from the given zone. Since there is a greater area for which regulation is anticipated, larger range areas may generate greater bias when range leases are unaccounted for in comparison to when there is no range area (i.e., the distance is zero because the zone and range boundaries coincide). Though these variables are similar, their correlation of approximately 0.5 implies that they are not perfect substitutes for one another.

We report coefficient estimates from estimating equation (3) using the two different measures of range size in Table 6. Panel A reports estimates from using the count of range leases as the size of a range, while Panel B reports estimates from using the maximum distance from a range lease.

Consider first the estimates in Panel A. The coefficient estimates for the zone variable,

Table 6: How the size of the bias from not accounting for anticipated rezoning is affected by the properties of potentially rezoned area

	(1)	(2)	(3)	(4)	(5)
Panel A: Interacting zone with total number of leases in range					
Zone	-0.2695 (0.3009)	-0.6242** (0.2152)	-0.2497 (0.1901)	-0.2656* (0.1353)	-0.3166** (0.1363)
× Lease count in range	0.0042 (0.0031)	0.0077*** (0.0024)	0.0017 (0.0016)	0.0001 (0.0013)	0.0004 (0.0014)
R^2	0.0051	0.1091	0.1539	0.5531	0.5539
Panel B: Interaction zone with maximum distance from within range to zone boundary					
Zone	-0.6707 (0.3968)	-0.6198 (0.4917)	-0.5550 (0.3224)	-0.4574* (0.2180)	-0.4587** (0.1971)
× Max. distance in range	0.0228 (0.0151)	0.0153 (0.0175)	0.0136 (0.0112)	0.0066 (0.0056)	0.0058 (0.0057)
R^2	0.0116	0.0992	0.1554	0.5534	0.5541
Quad. poly. in lat/lon	N	Y	Y	Y	Y
Zone fixed effects	N	N	Y	Y	Y
Owner-Year fixed effects	N	N	N	Y	Y
Lease controls	N	N	N	N	Y
Observations	2918	2918	2918	2918	2918

Notes: The dependant variable is the logarithm of price per hectare at the lease level. Robust standard errors, adjusted for clustering by caribou zone, in parentheses. *, **, and *** denote estimates different from zero at the 10%, 5%, and 1% significance levels.

in the first row, are all negative in sign, statistically significant, and larger in magnitude than the estimates reported in Table 5. The coefficient estimates for the interaction term, in the second row, are all positive. Together, the signs of the two sets of coefficient estimates indicate that the bias on the price effect for leases located in zones depends positively on the number of unaccounted for range leases. In particular, if there are no range leases associated with zones (which is the case for three zones), the coefficient estimate in column (5) is equal to -0.3166 , which is numerically similar to the baseline estimate of -0.3536 in column (5) of Table 2, where we do account for which leases are inside ranges. If the number of unaccounted for range leases is equal to the sample mean of range leases associated with a zone, 66, then the coefficient estimate for the effect of being in a zone is statistically significant at the 1% level and equal to -0.2890 , which is virtually identical to the estimate in column (5) of Table 5. If the total number of range leases is equal to the maximum we observe in our sample, 122, then the coefficient estimate of the effect of lying inside a zone is equal to -0.2696 and is statistically significant at the 5% level. Together, these estimates indicate that the greater the number of unaccounted for range leases, the greater is the bias on the estimated effect on existing regulation on lease prices.

The coefficient estimates using the maximum distance of a range lease to a given zone boundary, reported in Panel B, exhibit a similar pattern to those in Panel A. The coefficient estimates for the zone variable, in the first row, are negative, statistically significant, and larger in magnitude than the same-column estimates in Table 5. The coefficient estimates for the interaction term, in the second row, are all positive, indicating zones with geographically larger range areas lead to greater bias in the estimated effect of regulation when not accounting for which leases lie in those range areas. If the maximum distance of range leases associated with a given zone is equal to the minimum, about 250 meters, then the coefficient estimate for the effect on the price of a lease inside a zone in column (5) is equal to -0.4571 and is statistically significant at the 5% level. If the maximum distance is equal to the sample average, about 31 kilometers, the coefficient estimate is equal to -0.2777 , is statistically significant at the 1% level, and is virtually identical to the coefficient estimate in column (5) of Table 5. If the maximum distance is equal to the maximum observed in the sample, about 68 kilometers, then the coefficient estimate is equal to -0.0623 with a standard error of 0.2346. Altogether, these estimates imply that the bias from failing to account for which leases are inside ranges, thus anticipating future rezoning, depends on the area of the range those leases fall into.

Overall, we take this evidence in this section as supporting the prediction from the model

in Section 2 that not accounting for which leases may be subject to potential rezoning will lead to underestimates of the effect of existing regulation on land prices.

5 Heterogeneous effects of anticipated rezoning

The results reported in the previous section support the prediction that the expected cost of anticipated regulation is capitalized into the price of leases that may be rezoned in the future. In this section, we investigate whether the effect of anticipated regulation changes systematically as the likelihood of being rezoned plausibly varies.

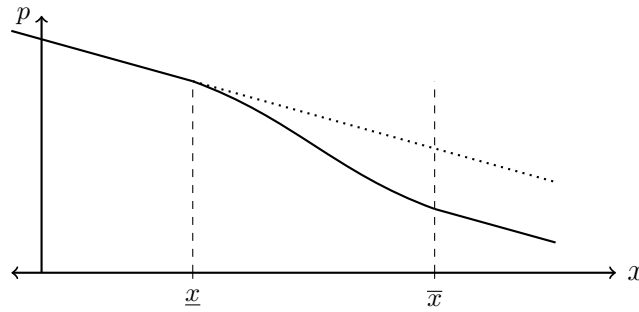


Figure 4: Land prices when rezoning is anticipated to happen gradually

As we described in Section 2, the expected costs of anticipated regulation are a combination of the likelihood of being rezoned and the cost of complying with the regulation. In Figure 1c, we depicted the likelihood of being rezoned, and thus the expected cost of anticipated regulation, as being constant over the unregulated area where rezoning is anticipated. However, it may be the case that zones are anticipated to undergo gradual expansion, so that parcels near the zone boundary \bar{x} face a greater likelihood of being regulated sooner than parcels farther away. In that case, the prices depicted in Figure 1c may instead look like they do in Figure 4. In the latter and for illustrative purposes, we assume that parcels lying just to the left of the zone boundary \bar{x} face a probability of being rezoned that is equal to 1, so that the expected cost of anticipated regulation is exactly equal to the cost of regulation; in this case, as depicted, there is no discontinuity in prices around \bar{x} . At the other end, \underline{x} , the probability is equal to 0, so expected costs are equal to 0 and there is also no discontinuity in prices there.²⁴ For parcels in between the two boundaries, the probability of rezoning

²⁴If the probability is not equal to 1 near \bar{x} , then there will exist a discontinuity in prices, though not likely equal to the cost parameter. Similarly, if the probability is not equal to 0 near \underline{x} , a price discontinuity will exist there.

increases with x , so that prices decrease at a faster rate toward \bar{x} than do the counterfactual prices under no regulation, depicted as the dotted line.

We can test the prediction depicted in Figure 4 by augmenting our baseline estimating equation (2), although the scope for precise estimation is less than ideal given we have only 320 observations of leases in ranges. In particular, we estimate an equation of the following form:

$$p_{izjt} = \alpha + \beta_1 \text{zone}_i + \tilde{\beta}_2 \text{range}_i + \tilde{\beta}_3 \text{range}_i \times \text{distance}_i + W_i \Gamma + f(x_i, y_i) + \lambda_z + \lambda_{jt} + \varepsilon_{izjt}, \quad (4)$$

where distance_i is a measure of distance to the nearest zone boundary for lease i . The effect of lying in the range area is equal to $\tilde{\beta}_2 + \tilde{\beta}_3 \text{distance}_i$. At the mean value for the distance measure, this effect should equal β_2 from equation (2).

We use three different measures of distance. The first is the distance of a lease, in kilometers, from within the range area to the nearest zone boundary. If range leases near zone boundaries face greater expected costs from anticipated rezoning than leases farther away, then we expect that $\tilde{\beta}_3 > 0$ with this distance measure.

Because the distance of range boundaries to zone boundaries differs by the zone and range (as can be seen from Figure 3), we create two alternative measures to account for those differences. The first of these two standardizes the distance from a range lease to the nearest zone boundary, so that the mean and standard deviation of this measure are 0 and 1, respectively, for all leases associated with a particular zone. The second measure takes the distance to the nearest zone and divides it by the sum of the distances of the lease to the nearest zone and range boundaries, so that this variable naturally lies between 0 and 1. Under both measures, the larger the value of the measure, the farther is a lease from the nearest zone boundary, in which case we expect that $\tilde{\beta}_3 > 0$ for both measures if expected costs of anticipated regulation are larger the nearer is a lease to a zone boundary.

We report the results from estimating equation (4) in Table 7. For convenience, we report estimates based on specifications using the full set of controls and fixed effects. The only difference in specification across the columns is in which of the three distance measures we use.

The specification in column (1) of Table 7 includes the interaction of the range variable with the distance, in kilometers, of a range lease to the nearest zone boundary. The estimate for coefficient $\tilde{\beta}_2$, -0.2687 , is statistically significant at the 10%, while the coefficient estimate for the interaction term is positive, though insignificant at conventional levels. Together, the signs of estimates imply that leases in range areas close to a zone boundary are subject

Table 7: Heterogeneous effects of anticipated rezoning depending on proximity to boundaries

	(1)	(2)	(3)
Zone	-0.3604*** (0.0850)	-0.3582*** (0.0834)	-0.3702*** (0.0844)
Range	-0.2687* (0.1513)	-0.1699 (0.1151)	-0.7291*** (0.1597)
×Distance to zone boundary	0.0096 (0.0092)		
×Std. distance to zone boundary		0.0705 (0.1458)	
×Zone distance ratio			0.7012*** (0.1841)
R^2	0.5545	0.5545	0.5551
Observations	2918	2917	2918

Notes: The dependant variable is the logarithm of price per hectare at the lease level. All columns report estimates based on specifications that include a quadratic polynomial in longitude and latitude, zone fixed effects, firm-year fixed effects, and lease-specific controls. Robust standard errors, adjusted for clustering by caribou zone, in parentheses. *, **, and *** denote estimates different from zero at the 10%, 5%, and 1% significance levels.

to greater price per hectare discounts than leases farther away. This finding is consistent with the prediction depicted in Figure 4 and is consistent with the notion that rezoning is anticipated to occur, based on historical practice, through gradual shifts in zone boundaries. In particular, for the range lease that is closest to a zone boundary – less than 20 meters from the boundary – the coefficient estimate of the effect is statistically significant at the 10% level and is equal to -0.2675 and implies a price per hectare discount of about 23.5% relative to unregulated leases. Given that our estimate for the effect of lying in a zone causes price per hectare discounts of more than 30%, this implies that range leases virtually adjacent to a zone boundary internalize expected costs that are smaller in magnitude than the costs capitalized into prices for regulated leases. The estimates imply that a range lease that is 5 kilometers from the boundary incurs a statistically significant 22% price per hectare discount relative to the prices for unregulated leases. The effect decreases in magnitude the farther is a lease from the boundary, and for leases that are greater than about 5 kilometers the price effect is statistically insignificant; for example, the negative price per hectare effect for a lease that is 20 kilometers from the nearest zone boundary, about one-standard deviation farther than the mean distance, is about 7.4% on average.

Column (2) reports estimates from a specification using the standardized distance to the nearest zone boundary as the distance measure. Similar to the estimates in the previous

column, the sign of the coefficient estimate for the interaction term is positive, implying that leases that are closer to zone boundaries are priced lower, all else being equal, than leases in the range that are farther away. The coefficient estimate for the range variable, -0.1699 , is similar to the baseline estimate reported in column (5) of Table 2 because the mean of the standardized distance measure is, by construction, equal to 0. A one-standard deviation decrease from the sample mean in the distance to the nearest zone boundary implies a price per hectare discount of 21.4% relative to unregulated leases, while a lease that is one-standard deviation farther than the sample mean from the nearest zone boundary incurs a price per hectare discount of 9.5% on average relative to unregulated leases.

In the final column, we report estimates from using the second alternative distance measure. The nature of this variable – it lies between 0 and 1 – results in an estimate of $\tilde{\beta}_2$, the coefficient for the zone variable, equal to -0.7291 , though the estimate given the mean value of the distance measure implies a price per hectare discount of about 17.7% on average relative to unregulated leases. Like the estimates in the previous columns, the positive coefficient estimate on the interaction term implies that leases in the range nearer to zone boundaries have prices per hectare that are lower than leases farther away in the range, all else being equal. For a lease that is closer to the nearest zone boundary by one standard deviation relative to the sample mean, the point estimate is equal to -0.3756 and is statistically significant at the 1% level; this estimate is virtually the same as the estimate describing the effect of being located inside a zone, reported in the first row. A one-standard deviation increase in the measure from the mean implies a price per hectare discount of only 1.3% on average relative to unregulated leases.

Altogether, the estimation results in this section support the prediction that the effect on the price for range leases increases the nearer is a lease to the zone boundary. This evidence is consistent with the notion that future regulation in the form of expanding zones are anticipated to occur through gradual boundary shifts, which was the historical practice in this context up until that time. However, it is noteworthy that the boundary changes in 2013 did not follow historical practice: instead of gradual expansion of the zones, the boundaries shifted dramatically so that every zone boundary coincided with every range boundary. In some cases, that meant a shift of more than 50 kilometers, which, based on the results reported in Table 7, was not anticipated to occur soon by the firms that purchased leases. Because firms' anticipation about potential future regulation relied on their likely belief that zones would expand gradually – as they always had been – it is not a surprise that they underestimated the likelihood that distant range leases would be rezoned relatively

soon.

6 The cost of existing and anticipated regulation

Based on our empirical strategy, we have identified the effect of existing and anticipated land-use regulation on auction prices for oil sands leases. Furthermore, the analysis in Section 2 suggests that we can interpret this effect as the present value (PV) cost of complying with, depending on a lease's location, existing or anticipated regulation over the lifetime of the lease. Given our estimation results indicate that the protection zones and the anticipation of their potential expansion cause negative effects on oil sands lease prices, this section puts those estimates into context by calculating the aggregate PV of existing and anticipated regulation for all leases in our sample.

To quantify the total cost of regulation, we need to compute lease prices under the counterfactual scenario where there is no regulation and then calculate the cost, after accounting for royalties and taxes, by subtracting observed prices from counterfactual prices. To see this, suppose the observed price for a lease is p . For a lease inside a zone, the effect of lying in the zone, based on our baseline estimate from column (5) of Table 2, decreases the price by 30% on average. The counterfactual price in the absence of regulation, which we denote by \tilde{p} , for a lease lying in a zone is therefore $\tilde{p} = p/(1 - 0.30)$. Similarly for a lease lying in a range, the baseline estimate from column (5) of Table 2 yields a price discount of 16% on average, so $\tilde{p} = p/(1 - 0.16)$. For unregulated leases, the absence of regulation has no effect on their prices, so $\tilde{p} = p$ for any lease located outside of zones and ranges. We perform this calculation for each lease in our sample, and calculate the price, as opposed to price per hectare, by accounting for the geographic area of the lease.

The difference between a lease's observed price and its counterfactual price describes only the loss of auction revenue from regulation. To calculate the PV cost of regulation, we need to account for corporate taxes and royalties on profits because in Alberta any costs associated with complying with environmental regulations are deductible from these taxes. For example, if the sum of corporate and royalty tax rates is denoted by α , where $0 < \alpha < 1$, a lease with lifetime real revenue R that is inside the zone, thereby incurring a cost of regulation, c , generates PV profits equal to $(1 - \alpha)(R - c)$. The PV profits equal the value of holding the lease and thus equal the price the lease commands in a competitive, first-price, sealed-bid auction. The difference in prices between a regulated and unregulated but otherwise identical lease is equal to $(1 - \alpha)c$. To calculate c , we then need to account for

the sum of royalty and corporate tax rates, α , applied to oil sands leases. At their lowest, the provincial and federal tax rates during this time are 10% and 15%, respectively. Royalty rates in Alberta depend on several different factors, and the minimum royalty rate for oil sands operations during this period is 25%, while the maximum rate is 40%; we use both rates to provide a range of cost estimates.

We depict four different categories of aggregate PV cost from the regulation in Figure 5. The height of the bars indicate total PV cost in millions of 2012 Canadian dollars, for the specified category, for all leases sold during 2003-2012. For each category, we depict the cost under the minimum royalty rate in dark grey, while the cost under the maximum rate is the lighter color.

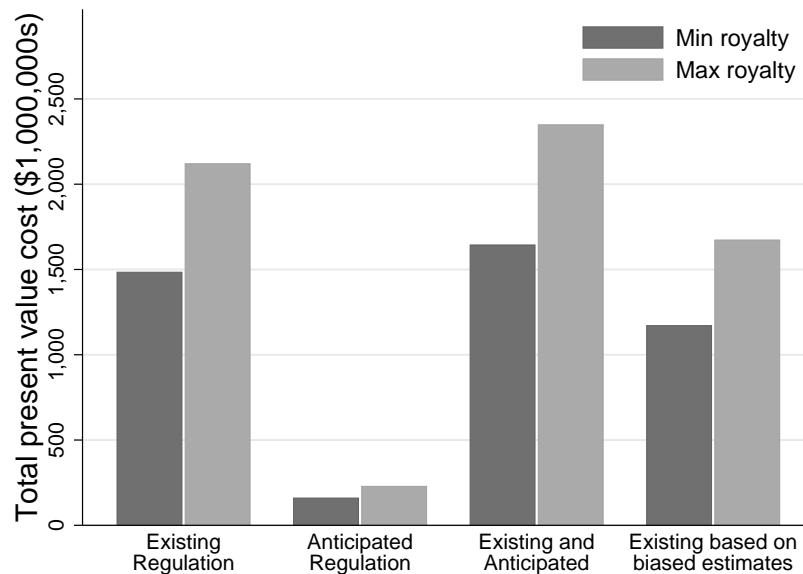


Figure 5: Total present value cost of regulation based on minimum and maximum royalty rates when anticipated rezoning is and is not accounted for in estimation

The first category is the PV cost of the regulation for all leases located in zones. Under the minimum royalty rate, the estimated total cost incurred for leases in the zones is \$1.484 billion. The estimated total cost under the maximum royalty is \$2.120 billion.

The second category is the cost imposed on the 320 leases that are outside zones – and thus are unregulated – but are located in ranges where regulation is not currently applied but where rezoning in the near future is plausibly anticipated. Based on the estimated price per hectare discount of 16%, the minimum PV cost imposed on these leases is \$160 million and the maximum cost estimate is \$228 million.

Because the regulation imposes costs on both leases that are regulated and leases that are unregulated but for which potential regulation in the future is anticipated, we calculate the total PV cost of the regulation as the sum of the costs to leases in the zones and the ranges. At a minimum, this cost is equal to \$1.645 billion, while at a maximum the total cost is equal to \$2.348 billion.

The cost of anticipated regulation accounts for nearly 10% of the total cost of regulation. Failing to account for anticipated regulation not only implicitly values those costs as being equal to \$0, but also, as we found in Section 4.3, leads to underestimates of the effect of protection zones. Based on the biased estimates in column (5) of Table 5, the price per hectare discount when omitting anticipation about rezoning is about 5 percentage points smaller than when anticipation is accounted for. The aggregate PV cost based on these biased estimates, depicted as the fourth category in Figure 5, is \$1.170 billion at a minimum and a maximum of \$1.673 billion. In comparison to the PV cost estimate for existing regulation when anticipated rezoning is accounted for in estimation, represented as the first category in Figure 5, these cost estimates understate the cost of existing regulation by about 22%. Relative to the total cost estimates which incorporate the cost of existing and anticipated regulation, the third category in Figure 5, the total cost estimates calculated from estimation that fails to account for anticipated regulation underestimates the total cost of regulation by 29%.

These estimates indicate that the prevailing approach to protecting endangered caribou in Alberta from the adverse effects of oil sands development is costly. Furthermore, all of the costs depicted in Figure 5 are borne entirely by the government: the costs from foregone auction revenues are incurred on the auction date, but the foregone royalty revenues are incurred during the production phase of each lease. And as Hervieux et al. (2013) find, caribou populations in Alberta are in continued decline and remain at ever greater risk of extirpation. Together with our finding that the regulation is costly and is incurred entirely by the government, this implies that the regulation is ineffective.

7 Conclusion

This paper examines the effects of existing and anticipated regulation on the value of land. The conventional approach to identifying the effect of regulation, regression discontinuity (RD), compares land values across administrative boundaries. In many settings, however, regulation boundaries can change over time, leading to potential identification issues. In

a simple model of land use regulation and potentially changing boundaries, we show that comparing land values around existing regulation boundaries underestimates the effect of regulation if forward-looking agents anticipate a change in the boundaries. We show that by accounting for such anticipation, one can identify not only the cost of existing regulation, but also the cost of anticipated regulation.

We apply this approach to auction prices for oil sands leases and zoning regulations protecting endangered species that are plausibly anticipated to expand in the Canadian province of Alberta. We find that the existing regulation causes lease prices to decrease by 30% per hectare on average, while anticipated rezoning causes the price per hectare of unregulated leases to decrease by 16% on average. This latter effect is most pronounced at existing regulation boundaries, where based on historical regulatory practice, rezoning is most likely to occur, and the effect dissipates the farther is a lease from those boundaries.

The results in this paper provide several lessons. First, our model shows that accounting for anticipation effects caused by potentially changing administrative boundaries is critical to identifying the effect of regulation on land values. Not accounting for anticipation underestimates the effect of regulation. This phenomenon extends beyond land use regulation to any context where administrative boundaries defining where policy is applied, providing useful variation to identify the effect of the policy, can change over time. Second, our empirical results indicate that anticipation effects matter in practice: anticipation about potential rezoning significantly decreases the price of unregulated oil sands leases, resulting in a loss of government revenue exceeding \$100 million. Furthermore, failing to account for anticipation effects underestimates the total cost of the regulation by nearly one-third. Given that we estimate the present value cost of the regulation to be in the billions of dollars, such a difference in cost estimates underscores the importance of accounting for anticipation when evaluating policy based on land values.

Our paper has broader lessons for policymakers. While delaying potential future regulation may have its benefits, the behavior by consumers or, in our case, firms caused by anticipating future regulation may come with considerable costs. In such a case, it is optimal for policymakers to weigh those costs, in addition to carefully considering how to signal future policy change, when designing public policy.

References

- Alberta Woodland Caribou Recovery Team (2005). *Alberta Woodland Caribou Recovery Plan 2004/5-2013/14*. Edmonton, AB: Alberta Sustainable Resource Development. Alberta Species At Risk Recovery Plan No. 4.
- Black, S. E. (1999). Do better schools matter? Parental valuation of elementary education. *Quarterly Journal of Economics* 114, 577–599.
- Bogart, W. T. and B. A. Cromwell (2000). How much is a neighborhood school worth? *Journal of Urban Economics* 47, 280–305.
- Bošković, B. and L. Nøstbakken (2016). The cost of endangered species protection: Evidence from auctions for natural resources. NHH Department of Economics working paper no. 6/2016.
- Boslett, A., T. Guilfoos, and C. Lang (2016). Valuation of expectations: A hedonic study of shale gas development and New York’s moratorium. *Journal of Environmental Economics and Management* 77, 14–30.
- Boutin, S., M. S. Boyce, M. Hebblewhite, D. Hervieux, K. H. Knopff, M. C. Latham, A. D. M. Latham, J. Nagy, D. Seip, and R. Serrouya (2012). Why are caribou declining in the oil sands? *Frontiers in Ecology and the Environment* 10(2), 65–67.
- Coglianese, J., L. W. Davis, L. Kilian, and J. H. Stock (2016). Anticipation, tax avoidance, and the price elasticity of gasoline demand. *Journal of Applied Econometrics*. Forthcoming.
- Dell, M. (2010). The persistent effects of Peru’s mining *Mita*. *Econometrica* 78, 1863–1903.
- Dzus, E. (2001). Status of the woodland caribou (*Rangifur tarandus caribou*) in Alberta. Technical report, Alberta Environment, Fisheries and Wildlife Division, and Alberta Conservation Association, Wildlife Status Report no. 30. Edmonton, AB.
- Environment Canada (2012). Recovery strategy for the woodland caribou (*Rangifur tarandus caribou*), Boreal population, in Canada. *Species at Risk Act Recovery Strategy Series*.
- ERCB (2013). ST98-2013: Alberta’s energy reserves 2012 and supply/demand outlook 2013-2022. Technical report, Alberta Energy Resources Conservation Board. <http://www.aer.ca/documents/sts/ST98/ST98-2013.pdf>.
- Government of Alberta (2009). Alberta’s oil and gas tenure. Ministry of Energy.
- Government of Alberta (2013). Enhanced approval process integrated standards and guidelines.

- Greenstone, M. and T. Gayer (2009). Quasi-experimental and experimental approaches to environmental economics. *Journal of Environmental Economics and Management* 57, 21–44.
- Hervieux, D., M. Hebblewhite, N. J. DeCesare, M. Russell, K. Smith, S. Robertson, and S. Boutin (2013). Widespread declines in woodland caribou (*Rangifer tarandus caribou*) continue in Alberta. *Canadian Journal of Zoology* 91, 872–882.
- International Energy Agency (2015). *World Energy Outlook*. OECD/IEA.
- Lee, D. S. and T. Lemieux (2010). Regression discontinuity designs in economics. *Journal of Economic Literature* 48, 281–355.
- Lueck, D. and J. A. Michael (2003). Preemptive habitat destruction under the Endangered Species Act. *Journal of Law and Economics* 46, 27–60.
- Malani, A. and J. Reif (2015). Interpreting pre-trends as anticipation: Impact on estimated treatment effects from tort reform. *Journal of Public Economics* 124, 1–17.
- Ries, J. and T. Somerville (2010). School quality and residential property values: Evidence from Vancouver rezoning. *Review of Economics and Statistics* 92, 928–944.
- Turner, M., A. Haughwout, and W. van der Klaauw (2014). Land use regulation and welfare. *Econometrica* 82, 1341–1403.
- Watkins, G. (1975). Competitive bidding and Alberta petroleum rents. *Journal of Industrial Economics* 23, 301–312.
- Watkins, G. and R. Kirby (1981). Bidding for petroleum leases: recent Canadian experience. *Energy Economics* 3, 182–186.
- Zabel, J. E. and R. Paterson (2006). The effects of critical habitat designation on housing supply: An analysis of California construction activity. *Journal of Regional Science* 46, 67–95.
- Zhou, J., D. P. McMillen, and J. F. McDonald (2008). Land values and the 1957 Comprehensive Amendment to the Chicago Zoning Ordinance. *Urban Studies* 45, 1647–1661.